

## RESEARCH OUTPUTS / RÉSULTATS DE RECHERCHE

### **Monoculture-based consumer-resource models predict species dominance in mixed batch cultures of dinoflagellates**

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# Harmful Algae

## Monoculture-based consumer-resource models predict species dominance in mixed batch cultures of dinoflagellates.

--Manuscript Draft--

<b>Manuscript Number:</b>	HARALG-D-19-00245R1
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<b>Abstract:</b>	<p>Global change will upset the frequency, scale and distribution of harmful algal blooms (HABs), but we are unable to predict future HAB occurrences due to our limited understanding of how physicochemical changes affect interspecific interactions between HAB and non-HAB species. Trait-based mechanistic modelling is an important tool to unravel such mechanisms and quantify the various direct and indirect interactions within systems. The present study explores whether MacArthur's consumer-resource model can describe resource competition between multiple HAB and non-HAB dinoflagellates. To this end, two batch culture experiments (294 cultures in total) with monocultures and mixed cultures of HAB (<i>Alexandrium minutum</i>, <i>Prorocentrum lima</i>, <i>Protoceratium reticulatum</i>) and non-HAB species (<i>Prorocentrum micans</i>, <i>Scrippsiella trochoidea</i>) were performed. Despite changes to the relative (the N:P ratio) and absolute nutrient availability (dilutions of L1 medium), <i>P. micans</i> continuously outcompeted all other species in mixed cultures. Consumer-resource modelling parameterized using monoculture growth correctly predicted this outcome (<math>R^2</math> between 0.80 and 0.95). Parameter estimates revealed that <i>P. micans</i> had a faster uptake of nitrogen when compared to its competitors, but did not differ in resource efficiency and natural mortality rate. Yet, while the model accurately predicted community dynamics during the growth phase, it was not able to predict their dynamics beyond the point of quiescence. Overall, consumer-resource modelling was shown to differentiate the roles of resource assimilation, resource efficiency, and natural mortality rates in these common experiments with minimal data requirements.</p>



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Dear Dr. Kudela,

On behalf of my co-authors, I am pleased to deliver the revised manuscript of our original research article - entitled "**Monoculture-based consumer-resource models predict species dominance in mixed batch cultures of dinoflagellates.**" – to be considered for publication in *Harmful Algae*.

In this manuscript, we explore the potential benefits of consumer-resource models (CRMs) as a basis for trait-based modelling of community dynamics of dinoflagellates in multispecies cultures. To this end, we adapt MacArthur's CRM – a largely forgotten adaptation of the Lotka-Volterra equations that enjoyed limited success in theoretical ecology – for general use in HAB research, and assess its capabilities by fitting it to the growth data of 294 single and mixed batch cultures. Trait-based modelling is rarely used on lab cultures, but it could advance our understanding of the relative importance of various interspecific interactions (incl. resource competition, allelopathy and grazer deterrence). Here we show that CRMs are well suited to understand and predict the resource competition between dinoflagellates in mixed batch cultures (the most common of culture methods). They are easy to use, require minimal data, and provide key insights into the importance of nutrient uptake, conversion efficiencies and maintenance requirements when comparing various (harmful) algae. We then discuss various novel model improvements that may extend this trait-based approach to studies on allelopathic interactions as well as in situ predictions.

The manuscript was originally submitted to *Harmful Algae* in Nov. 2019 and came back with "Major Revision" in Feb. 2020. Both reviewers agreed that the manuscript was interesting and found the proposed model to be potentially useful to the HAB community, but raised several points that needed addressing. We have copied these in our Response to the Reviewers.

We remain convinced that CRMs hold a lot of (mostly unexplored) potential for HAB research and, hence, believe that *Harmful Algae* would be the most appropriate journal for our manuscript. We have no conflicts of interest to disclose. The manuscript is an original work that has not been published, nor is it under consideration for publication elsewhere.

Thank you for your time.

Sincerely,

Maarten De Rijcke; Jan Baert  
Natacha Brion; Michiel Vandegehuchte  
Frederik De Laender; Colin Janssen

*Dear Dr. Kudela,*

*Thank you for considering our manuscript titled "Monoculture-based consumer-resource models predict species dominance in mixed batch cultures of dinoflagellates." (HARALG-D-19-00245) for publication in Harmful Algae. Below you may find the appraisals of both reviewers and point-by-point responses to specific comments made by each. Enclosed, you will find our manuscript that was adapted accordingly. We would be glad to respond to any further questions and comments that may arise and look forward to hearing from you.*

*Yours sincerely,  
Maarten De Rijcke (on behalf of all co-authors).*

### Reviewer 1 – General appraisal

The manuscript entitled: "Monoculture-based consumer-resource models predict species dominance in mixed batch cultures of dinoflagellates." is an innovative research paper which tackles the difficult question of predicting the HAB biomass. The authors used a mechanistic approach to the question by elegantly improving MacArthur's (1970) consumer-resource model with the parameters estimated by the results of the mono-cultures experiment. The model was hereafter used to predict the outcome of the mixed-cultured experiments from which it was partly successful. The procedure was repeated, first, at different N:P ratio treatment, and in second at different N:P ratio and absolute concentration. They did a tremendous experimental work to get the required parameters and to test their newly improved model. I believe to be a really inventive and elegant way forward which will sure gather a lot of interest within the HAB scientist community and even beyond. However, it is clear that the paper mostly tried to promote the model rather than the experimental set-ups, which I believe deserve more credits.

From the materials and methods to the end, I had the sense that the experimental part of each section (M&M, results and discussion) did not have as much attention, scrutiny and clarity than the modelling part. For instance, the M&M section on the first and second experiments (2.2 and 2.3) need to be more fluid. Also, despite mentioning the different statistical analyses, they are not explanation on what they use for. What do they compare with each statistical analysis and why? On the other hand, the "2.5 Development of a community model" section was easy to understand even for me who had never work with such model. As there is a back-and-forth between the experiments and the model I think it would help to have a flow chart illustrating the method design and would greatly help the readers to understand as it is rather complex.

I also had a hard time reading the results section, which to my opinion needs to be rewritten with the appropriate figure or table to illustrate the text. As we don't know what they have compared with each statistical method, we end up with a raw string of p value. Also, it needs more references to the appropriate figure and tables. Fig.3 is not referenced in the results section but only in the discussion which I think it is also useful beforehand. I think the flowchart in M&M could help clarify the needs for the different results and what they are used for.

The discussion suffers the same problem as the M&M section. Even if the authors mention that "The results of our experiments should not be viewed as ecological stoichiometry research" I believe that they should still talk about the problem they encountered during their lab experiments. Why didn't they manage to get the P.lima carrying capacity? Can they provide an explanation? Also, why do they think there was no cell growth in CF1 and CF0.1? What is their explanation? How would they've done it differently? As they provided new insight upon their model, I think they should make the same effort with the experimental section.

After my review I think that this work should be published after major revision as, despite the problem mentioned above, it has a great potential.

*We would like to thank Reviewer 1 for the detailed comments that have helped us improve the manuscript. We appreciate the feedback that the model section was easy to understand as it has indeed received most scrutiny to that end. With the help of the comments of Reviewer 1, which we copied below, we hope to have improved the manuscript to similar effect.*

## Reviewer 1 – Specific comments

### **Introduction**

I. 31 replace "upset" by "disturb"

A: We replaced “upset” by “disturb” as suggested (l. 31).

I. 71 "the succession of groups of phytoplankton" or "the phytoplankton groups succession".

A: We changed “the succession of phytoplankton groups” to “the succession of groups of phytoplankton” as suggested (l. 72).

I. 79 "appeared" or "seemed to be linked"

A: We replaced “can be linked” by “seem to be linked” to accommodate this remark (l. 80).

I. 83 "altering every biological interactions"

A: We used “every biological interaction” as suggested (l. 86).

I. 88 "In order to coop"

A: We replaced “to coop” with “In order to cope” to incorporate the suggestions of both reviewers (l. 91).

I. 91 "Their ability"

&

II. 91 - 95 a bit confusing and the end of the sentence is not clear

A: We agree that the sentence could have been more concise. We have combined and rephrased this section as follows:

“Toxins, grazer deterrents and allelochemicals - i.e. exudates that cause nutrient leakage, inhibit photosynthesis, arrest the cell-cycle, or affect other enzymes of competing algae (Granéli and Hansen, 2006; Legrand et al., 2003; Reigosa et al., 1999) - reduce the long-term extinction risk of toxic algae and may help maintain toxic blooms (Granéli et al., 2008a; Ianora et al., 2011; Smayda, 1997; Smayda, 2008; Turner, 2006; Xu and Kiørboe, 2018).” (ll. 95-99).

II. 100-104. "Because of the variable .... during bloom initiation." The sentence is unclear and too long.

A: We have rephrased and split the sentence to improve this section. It now reads:

“Toxic effects are variable or inducible in nature (e.g. Dam and Haley, 2011; Poulin et al., 2018), mostly occur at bloom-level densities (Jonsson et al., 2009), and can have a high individual cost for toxic species while providing collective benefits to others (Driscoll et al., 2016; Flynn, 2008a). For these reasons, there is doubt that these chemical interactions play a crucial role during bloom initiation.” (ll. 102-106).

I. 104-105 I suggest "More recently, Blossom et al. (2019)..."  
&

I. 106 "low cell concentration. Moreover, they also suggested"

A: The sentences were changed according to these suggestions (II. 106-109).

II. 107-109 "The processes..., so to better understand the non-deterministic"

A: The proposed sentence felt disconnected from the prior text due to the changes detailed above. We have changed the sentence to "Overall, these studies demonstrate that the processes behind allelopathy need to be unravelled further to understand the non-deterministic nature of HABs during windows of opportunity." instead (II. 109-111).

I. 118-121 Split the sentence

A: We split the sentence as well the subsequent long sentence. The section now reads:

"This issue is often addressed by use of the dilution method. By increasing the number of target cells relative to a constant density of an allelopathic species, the amount of allelochemicals per target cell decreases. This should lead to increased growth of the target species. If the growth rate remains constant or decreases due to an increase in competition, allelopathy is considered to be absent (Weidenhamer, 2006)." (II. 120-124).

I. 124 "(1)..." not clear

A: Poulin et al. (2018) have shown that the allelopathic effects of *K. brevis* on *A. glacialis* varies from strongly inhibitory to strongly stimulatory between strains. When using the dilution method, we would surmise that decreasing growth rates at higher densities of target species are caused by increased intraspecific competition rather than stimulatory allelopathic interactions. To avoid discussing whether allelopathy should be reserved for negative effects (*sensu stricto* definition) or should include all chemical interactions between cells, we propose to rephrase the sentence.

We have added "between strains" to the sentence and replaced "voiding the null hypothesis" with "obscuring the interpretation of the test" (I. 127).

I. 127 replace "species-species" by "interspecific"

A: Replaced as suggested. (I. 129)

I. 128 remove "aquatic" unless terrestrial phytoplankton exist?

A: The word "aquatic" was removed from the sentence (I. 130).

I. 140 "introduced the resource utilization functions" ?

A: We added "resource" to the sentence as suggested (I. 141).

## **Material and Methods**

I. 168 How did you determine the exponential growth phase?

A: Weekly cell counts were performed on stock cultures. We added the following sentence to the manuscript to clarify this:

"The growth of stock cultures was monitored through weekly cell counts using a Sedgewick-Rafter counting chamber and a Kyowa Optical Biolux-2 light microscope. Both experiments used cells taken from stock cultures that were growing exponentially." (ll. 172-174)

I. 171 don't start your sentences with "To". Turn the sentence around, you must

A: We have restructured the sentence as suggested. (ll. 177-178)

I. 173 I suggest "Ten algal growth media were prepared so to have ten unique nitrogen-to-phosphorus ratios",

A: We have rephrased the sentence as suggested. (l. 179)

II. 174 -179 The description of your medium preparation can be simplified. The L1 medium is made of ... NO and ... PO corresponding to a N:P of 24. By only adjusting the NO concentration.... (all the NO concentration with all the N:P ratios).

A: We have shortened the corresponding paragraph (ll. 181-183) as follows:

"By only adjusting the  $\text{NO}_3^-$  concentrations, ten growth media with different nitrogen-to-phosphorus ratios were prepared: preparations of 294, 368, 441, 478, 515, 551, 588, 662, 735 or 882  $\mu\text{M}$   $\text{NO}_3^-$  corresponded to a N:P of 8, 10, 12, 13, 14, 15, 16, 18, 20 or 24, respectively."

I. 180 replace "in all media" by "in each medium"

A: We replaced "in all media" by "in each medium" as suggested (l. 188).

I. 182 so how many monocultures, mixed culture and in total you had? Show off a bit more the extend of your work

A: We set up 120 monocultures and 30 mixed cultures or 50 treatments. This is now included in the text (ll. 191-192)

I. 184 "a 1 ml sample"

A: This was changed as suggested. (l. 194)

II. 194 Change the sentence, you must

A: We have rearranged the sentence. Lines 203-204 now read:

"A second experiment was performed to examine whether the interspecific competition between dinoflagellates in batch cultures is affected by larger differences in macronutrient availability."

II. 196 It is not like experiment 1 as you had 10 conditions and in the second you have 12

A: We have removed the reference to experiment 1 as suggested. (l. 205)

II. 197-206 Again I think the explanation of the experimental set up can be simplified. If I haven't misunderstood. L1 with alter N:P ratio (No concentration and respective ratio). Each altered N:P ratio medium were submitted to four different dilution or concentration factors (CFs), so to have 100%,... In the end, 12 unique medium were made and used for monoculture of .... and a mixed culture with triplicates each time. So in total you have 144 cultures including 108 monocultures and 36 mixed cultures. Is that correct? If it is, well done for your job. Maybe a table or an illustration can help the reader to understand your setup

A: The reviewer is right: we made 144 cultures spread across 12 unique media. Taking inspiration from the previous comment on the design of experiment 1, we simplified the description of the setup and summarized the design of experiment 2. The corresponding sections now read:

“Twelve unique algal growth media were made based on regular L1 growth medium. Media were first prepared with 294, 588 or 882  $\mu\text{M}$   $\text{NO}_3^-$  to obtain three N:P ratios (8, 16 and 24). All other L1 components ( $\text{PO}_4^{3-}$ , vitamins and trace elements) were added at the regular dose. Each medium was subsequently diluted by a factor 1, 10, 100 or 1000 to obtain media with 100%, 10%, 1% or 0.1% volume fractions of L1 medium vs. Instant Ocean™ artificial seawater.” (ll. 205-209)

“Every treatment was replicated three times, resulting in 144 cultures (108 monocultures and 36 mixed cultures).” (ll. 214-215)

Why not using the same nutrient analysis for experiment 1 and 2?

A: The Analytical, Environmental and Geo-Chemistry research group of the University of Brussels owns the QuAAtro nutrient analyser. We did not have an active collaboration at the time of experiment 1, so we had to resort to a technique that could be performed at our lab.

I. 221 log-logistic? Logarithmic?

A: We used logistic growth models based on the Verhulst equation to describe the growth of our populations. In the earlier version of the manuscript, we incorrectly referred to log-logistic models and even logarithmic models, which has now been corrected throughout the manuscript.

I. 222 Kruskal Wallis and DMC references are missing

&

I. 225 Linear regression Ref?

A: We have added appropriate references for each method to the manuscript (ref. next comment).

I. 222-224 What are you going to compare? Between species? Between treatments? Both?

A: We used pairwise testing to compare the performance of each species to their growth in mixed cultures. Multiple group comparisons were used to compare growth rates and carrying capacities between species as well as to detect differences between treatments within each species. Linear regressions were used to detect linear responses to nutrient stoichiometry. We have rephrased the corresponding section to clarify the use of each test as follows:

“Multiple group comparisons by means of Kruskal Wallis (KW) tests (Kruskal and Wallis, 1952) were used to compare growth parameters ( $\mu$  and K) between treatments (N:P) and species. Pairwise comparisons using Dunn’s multiple comparison (DMC) test (Dunn, 1964) were made to investigate the effects of treatments (CF, mono vs. mixed) on the growth of each species. Linear regression models (LM) were used to detect linear responses to nutrient stoichiometry as described by Wilkinson and Rogers (1973).” (ll. 231-236).

I. 237 "Any excess of prey captured was converted into... into grams of Xi. "

A: We have replaced “any excess prey” by “any excess of prey” as suggested (l. 248).

I. 255 -256 "because nitrogen concentration vary the most in our experiments."

A: This sentence was rephrased in accordance with a comment of reviewer 2. (ll. 266-267)

II. 272-274 split the sentence, annealing algorithm

A: We have split the sentence at annealing algorithm as suggested. (II. 283-286).

I. 275 MCMC ref?

A: We have included Hastings (1970) as the appropriate reference as suggested (I. 286).

I. 276 - 277 Turn the sentence, you must

A: We have rearranged the sentence as suggested (I. 287-288).

## Results

I. 289 Why 28? Should it not be in the method section?

A: There was no reason beyond the practical. Due to technical problems at the lab, only 1 climate room was available. As all our mixed cultures had reached stationary growth / species dominance, we terminated exp. 1 so that other experiments at a different temperature could take place.

I. 289-301 Where is the carrying capacity of *P. lima*? If you don't have it you should explain why. p should be written in italic capital (*P*)

A: *P. lima* was still growing exponentially by the end of the experiment, so we were unable to determine and report its carrying capacity. We changed the suggested lines emphasize that this species was still growing. The manuscript now states:

“Logistic growth models were used to determine the monoculture growth rates for all species except *P. lima*, which was still growing exponentially at the end of the experiment. Exponential growth models were used to determine the growth rates of *P. lima* instead (Supporting figures SF1-4).” (II. 300-303).

We have replaced “p” with “*P*” for all statistics as suggested.

I. 302-311 "lost over half of its carrying capacity" where is the information? I also think that there is a figure or a table missing because I don't know what you are describing. The figure 12 in your supporting information (which should S6) has 20 graphs so which one are you describing?

A: The average carrying capacity of monocultures of *P. micans* was reported in the first paragraph of section 3.1. We now reiterate the monoculture value to help readers:

“On average, *P. micans* lost over half of its carrying capacity to competitors: its average carrying capacity decreased from  $5.5 \pm 1.4 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$  in monocultures to  $2.1 \pm 0.4 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$  in mixed cultures.” (II. 316-318).

The statement related to SF12 refers to the declining nutrient concentrations. These can be found in the right graph for each treatment, which was only apparent in the figure caption. We added legends to each plot within the figure. In addition, we removed the asynchronous referencing of supporting figures by modifying their order and fixing their references, making sure that they are now uniformly called Supporting Figure or SF. The modified section now reads:

“Nitrogen and phosphorus concentrations from the first experiment can be found in supporting figures SF6-10. In mixed cultures, nutrients were depleted in all but the highest N:P ratio by day 14 (Fig. SF10).” (II. 323-325)

I. 314 So CF 0.1 and 1 did not work out am I right? Need to write it

A: We observed between 1 and 3 cell divisions (population doublings) before the growth stopped. We added this observation to the text. Section 3.2 now starts with:

“The second experiment lasted 56 days, but the lowest concentration factors (CF0.1 and CF1) did not support prolonged growth. We observed between 1 (the lowest belonging to *P. reticulatum*) and 3 (found for *A. minutum*) population doublings before growth stalled. These treatments were no longer sampled after 39 days.” (ll. 328-331)

I. 317 remove "again"

A: We have removed “again” from line 333 as suggested.

I. 320-321 rewrite the sentence

A: We have rephrased the sentence. It now states:

“The N:P ratio did have a significant (LM  $P < 0.01$ ) positive effect on the carrying capacities of the three dinoflagellates at both CF10 and CF100.” (l. 335-337).

I. 322 remove "again"

A: We have removed “again” from line 338 as suggested.

II. 323-324 sentence not clear

A: We have split and rephrased the sentence:

“The growth rates of each species were determined by logistic growth models for CF10 and CF100. No significant differences were found between the growth rates of monocultures and mixed cultures for any of the three species (KW  $P > 0.05$ ).” (ll. 338-341)

I. 351 Isn't there something missing between the brackets?

A: The bracket was complete but too brief. We found that increased growth rates are significantly linked to higher nutrient uptake rates in both experiments through linear regression models. We changed the bracket to “(LM:  $P < 0.001$  for exp. 1; LM  $P < 0.01$  for exp. 2)” to reflect this (l. 369).

## Discussion

I. 361 "While many studies have" which one have you read?

A: There are numerous papers that investigate the effect of one or more parameters on the growth HAB species available in literature. We have added references to examples from the references that were already present in the study in lines 379-382. These are Chang and McClean (1997), Cooper et al. (2016), Gallardo Rodríguez et al. (2009), Guerrini et al. (2007), Ignatiades et al. (2007), John and Flynn (2000), Nascimento et al. (2005), Peperzak (2003), Sala-Pérez et al. (2016), Varkitzi et al. (2010), Wang et al. (2014), Zhengbin et al. (2006)

I. 363 "only a few have" where are the ref?

A: We now refer to Ji et al. (2011), Li et al. (2012), Poulin et al. (2018), Riegman et al. (1996) and Wang and Tang (2008) as examples. All these references were part of the original reference list. (ll. 383-384)

II. 378 -379 "the data shown" what table and/or graph?

A: Growth rates of this study can be found in section 3.1 and Table 2. We included a reference to these sections (l. 400).

I. 379 Modification of the macro(nutrient) concentration in the growth media

A: We have rephrased the sentence as suggested. (line 401).

I. 394 I suggest "the dinoflagellates community structure" instead of hierarchy

A: We replaced "the hierarchy of dinoflagellates" with "the dinoflagellates' community structure" as suggested. (line 417).

I. 398 CRM you might want to introduce it as consumer-resource model I presume?

A: The reviewer is right to point out that the abbreviation CRM was not introduced in the discussion (only in the introduction). It is now written "consumer-resource model (CRM)" in full (line 421).

II. 397-400 Split the sentence

A: We have split the sentence as requested. The section now reads:

"According to the mean parameter estimates of our consumer-resource model (CRM), the success of *P. micans* should be attributed to its ability to capture resources rather than a high resource efficiency or low natural mortality rates. The uptake probability of both nitrogen and phosphorus of *P. micans* were (among) the highest observed." (lines 420-424).

II. 401-402 replace by "in the two experiments"

A: We have replaced "during both the experiments" by "in the two experiments" as suggested. (line 419).

II. 400-404 split the sentence

A: We have split the sentence as suggested. The section now reads:

"All pelagic dinoflagellates grew at roughly the same rate relative to their monocultures in the first days of the two experiments. By sequestering nitrogen and phosphorus more rapidly, thereby denying its competitors access to these nutrients, *P. micans* was eventually able to outgrow all other species in mixed cultures"

I. 405 "in mixed cultures" instead of "in competition"

A: We replaced "in competition" by "in mixed cultures" as suggested. (line 428).

I. 410 "Luxury consumption" definition? Ref?

&

II. 410-412 rewrite the sentence for better clarity

A: The manuscript now references the original paper that coined the term "luxury consumption" as well as a recent review that covers its potential as a functional trait. We have also rewritten the next sentence as suggested. The entire section now reads as follows:

“Another unknown is whether the success of *P. micans* can be attributed to “luxury consumption”. The rapid acquisition and storage of excess nutrients may be used to pre-emptively reduce the availability of resources for competing species (Droop, 1973; de Mazancourt & Schwartz, 2012). This trait has not been studied in *P. micans* to our knowledge, but its carrying capacity is known to positively correlate with nitrogen concentrations (Zhengbin et al., 2006; Zheng-fang et al., 1995). Similar results were found in this study.” (ll. 432-437)

ll. 422-425 split the sentence

A: We have split the sentence as suggested. The section now reads:

“We set out to determine the efficacy of consumer resource modelling. Starting with the simplest setup available, which is the batch culture, we found that CRM’s could be used to predict species dominance resulting from interspecific competition between dinoflagellates in mixed cultures.” (ll. 447-450).

I. 460 "However, as shown here" which graph table?

A: The population decline can be observed in the density data of both experiments, as shown in supporting figures SF5 and SF12. We have added a reference to these figures in the statement. (line 484)

## Reviewer 2 – General appraisal

This study combines batch-culture experiments testing the growth of five different dinoflagellates in monoculture and in mixture under different nutrient regimes (varying N:P concentrations and ratios) with a consumer-resource model (CRM). This model was parameterized from monoculture growth and was used to predict the outcome of competition in species mixtures. Overall, *Prorocentrum micans* outcompeted all other species in mixed cultures irrespective of nutrient regime, apparently based on the dinoflagellate's faster uptake of nitrogen. The CRM correctly predicted this outcome, at least during the growth phase of the dinoflagellates. The authors claim that CRMs provide a useful tool to predict dominance of HAB versus non-HAB species and that their model may improve our understanding of HAB dynamics.

Overall, this manuscript yields interesting information and a potentially useful model approach for investigating competition of phytoplankton including HAB dinoflagellates. However, the study does not live up to its promises: The introduction strongly focuses on allelopathy and direct interactions of HAB species with competing non-HAB species. The authors distinguish between toxic and non-toxic dinoflagellates in their experiments at first, misleading the readers to expect a study explicitly testing for competitive interactions between allelopathic HAB and non-HAB species. However, neither the experiments nor the model were designed to explicitly investigate competition between HAB and non-HAB species, or the effect of allelopathy in competitive interactions. Rather, the authors seem to use a random set of dinoflagellates, some of which may also produce toxins and all of which have been shown to produce allelochemicals (which does not become clear until the end of the discussion). Allelopathic interactions were neither measured for the species, nor were they included in the model; therefore, it is not known whether allelopathy played any role at all in the experiments.

The authors should make very clear from the beginning on that this study investigates competition of dinoflagellates in general without taking direct interactions such as allelopathy into account, and that the model can be used for all competing phytoplankton species, while it is not specifically designed to investigate interactions of HAB versus non-HAB species.

As allelopathy or HAB versus non-HAB species do not play any role in either the experiments or the model, the introduction and parts of the discussion need to be completely rewritten, focusing on nutrient competition, stoichiometry and traits of the dinoflagellates used in this study. The major point in the discussion should be that a model that is solely based on nutrient concentrations, uptake probability and conversion efficiency can predict the competitive outcome of dinoflagellates, at least in the growth phase, while direct interactions may play a larger role at high bloom concentrations, which could be included in the hybrid model that the authors propose at the end of the discussion.

I recommend publication of this manuscript after a major revision setting the study into the right context, clearly motivating its intention and emphasizing its potential relevance. However, I leave it to the editor to decide whether this general competition study is suitable for publication in Harmful Algae.

*We would like to thank Reviewer 2 for his/her valuable input that helped us improve our work. We agree with the reviewer that our experiments do not adequately address HAB vs. non-HAB competition and have removed most references to HAB vs. non-HAB and toxic vs. non-toxic from the manuscript. The main conclusion of this pilot study is that CRMs provide a valuable basis for trait-based modelling of species interactions in mixed cultures. We believe that the use of mixed cultures coupled to improved CRMs (ref. discussion) can expand our understanding of allelopathic interactions. Lessons should be drawn from this study to design experiments to that end. For this reason, we introduce allelopathy in the introduction and conclude on allelopathy in the discussion.*

*Some of the reviewers' comments are related to our geographical bias. The research presented here was part of a larger project that aimed to investigate the present and future risk of HABs in the Belgian Part of the North Sea. The seemingly random set of dinoflagellates are all species occurring within our EEZ. Likewise, the N:P ratios and light treatment we used are based off of in situ observations. We provided further clarification below.*

## Reviewer 2 – Specific comments

### **Abstract & Highlights**

The information that is provided in the highlights should also be included in the abstract

A: We have modified the highlights to adhere to the abstract.

Please clarify that the experiments were performed with multispecies mixtures (as opposed to 2-species mixtures)

A: I. 38 now refers to “multispecies cultures” instead of “mixed cultures”

The authors should clarify the most important findings of their study (see also general comment)

A: The abstract now contains our main finding that CRMs are a potentially valuable basis for trait-based modelling that needs further development (II. 49-50)

### **Introduction**

I. 68: please add an "a" or "the" before "...major goal...."

A: I. 69 was changed as suggested (“a major goal”).

Line 80 - 82: please add that in addition to similar nutrient requirements and uptake kinetics to non-HAB species, there is a huge variability within and among different HAB groups in these parameters, hampering our predictive capability of where and when HABs will occur

A: We have rephrased II. 80-82 to include reference to the natural variability within species. The section now reads as follows:

“While both chronic and episodic eutrophication seem to be linked to HABs, there is no clear evidence that nutrients promote HABs by themselves. Nutrient uptake kinetics and resource preferences vary greatly within and between HAB species (Glibert and Burkholder, 2006) and cannot be distinguished from those of closely related non-HAB species (Anderson et al., 2002; Heisler et al., 2008; Wells et al., 2015). This hampers our ability to predict where and when HABs will occur based on resource abundance alone” (II. 80-85).

Line 88: please exchange "coop" by "cope"

A: This was changed as suggested (II. 90).

Line 90: allelopathy is not necessarily toxin-mediated; e.g. for Alexandrium it has been shown that allelopathic substances are not related to the production of PSP toxins (Tillmann & John 2002, MEPS 230)

A: We agree with the reviewer, that is why the original sentence listed allelochemicals, toxins and grazers deterrents as separate entities, but the sentence was unclear (see also reviewer 1). We have rephrased the sentence. The manuscript now reads:

“Toxins, grazer deterrents and allelochemicals - i.e. exudates that cause nutrient leakage, inhibit photosynthesis, arrest the cell-cycle, or affect other enzymes of competing algae (Granéli and Hansen, 2006; Legrand et al., 2003; Reigosa et al., 1999) - reduce the long-term extinction risk of toxic algae and may help maintain toxic blooms (Granéli et al., 2008a; Ianora et al., 2011; Smayda, 1997; Smayda, 2008; Turner, 2006; Xu and Kjørboe, 2018).” (II. 95-99).

Line 114 - 115: not necessarily, this is very dependent on the HAB species

A: The reviewer is right to point out that toxicity is not necessarily increased by direct contact. We have rephrased the sentence to be more cautious:

“The toxicity of intact cells can sometimes be increased by or be dependent on direct contact with targets (Driscoll et al., 2016).” (ll. 116-117).

Overall, the authors set a strong focus on allelopathy and on how to test for allelopathic interactions; however, neither in their experimental design nor in their model they account for allelopathic interactions. Therefore, the introduction and the actual study presented in Methods and Results seem a bit decoupled

A: As stated above, we believe that multispecies cultures coupled to improved CRMs can improve our understanding of allelopathic interactions. For this reason, we introduce allelopathy in the introduction and conclude on allelopathy in the discussion.

The authors should introduce the dinoflagellate species that they use in their experiment in terms of toxins, allelopathic substances, and especially potential nutrient requirements (have blooms been related to nutrient conditions) as experiments and models are designed to test nutrient effects on competition

A: A detailed introduction on the allelopathic and toxic properties of each species used may create false expectations from the reader. As the reviewer rightfully points out, we did not measure these chemicals during our study. We never measured the resource preferences and uptake kinetics for various N-sources of the strains that we used in the study either. As the reviewer pointed out, these can vary strongly within species. The information also has little bearing on the experiment, as the L1 medium only offers nitrate and phosphate as standardized nutrients.

The authors should motivate why they test different nutrient concentrations and ratios in their experiments - what role does nutrient stoichiometry play for algal competition, how do changes in nutrient ratios effect HAB species etc.

A: As explained in the discussion, our experiments are not suited to study ecological stoichiometry so we want to avoid creating false expectations in that regard. Generally speaking, we only used different nutrient concentrations (be it CF or N:P ratios) to introduce some variability for the model to predict and to try and upset the species dominance in mixed cultures. As the N:P ratios are not controlled throughout the experiment, they will have changed differentially during the experiment depending on the uptake kinetics of each species. Nutrient concentrations should also have been lower / limiting to study the effect of nutrient stoichiometry.

The authors distinguish between HAB and non-HAB species - however, this distinction does not play any role for further aspects of their study, neither for the experiments nor for the model, as the competition of all dinoflagellates in a multi-species mixture are investigated

A: The distinction between HAB and non-HAB has been removed from the manuscript.

I. 151: please specify that dinoflagellates were tested in multispecies mixtures containing "HAB" and "non-HAB" species, and not explicitly testing HAB versus non-HAB species in 2-species mixtures

A: We now refer to multispecies cultures throughout the manuscript and have removed references to HAB and non-HAB dinoflagellates.

## Materials and Methods

On what basis were the different N:P concentrations and ratios prepared? What was the intention of such a high resolution of N:P ratios?

A: The research presented here was part of a larger project that aimed to investigate the present and future risk of HABs in the Belgian Part of the North Sea. The N:P ratios used here are centred around 14, which is the mean N:P ratio found between June and September (years 2013-2019) in the marine area designated for aquaculture in Belgium. On an annual basis, the mean N:P ratio for all monitoring stations within the Belgian Part of the North Sea is 22. We have added a short reference to the M&M section to clarify that our nutrient ratios were inspired by our own location:

“The range of N:P ratios that was used was based on N:P ratios that were observed in the Belgian part of the North Sea between 2013 and 2019 (Mortelmans et al., 2019). On average, the Belgian EEZ has a mean N:P ratio of 22 while the average N:P ratio in a local shellfish area is around 14 during summer.” (Il. 184-187).

The L1 dilutions were inspired by the work of others, as referenced in the discussion (Il. 401-404).

Il. 181-182: why did the authors chose an additive instead of a substitutive design? Mixed cultures started with a much higher algal biomass than monocultures - are monocultures and mixtures comparable in terms of intraspecific versus interspecific competition?

A: We used an additive design to avoid human error while setting up so many cultures, but both designs would have worked as the consumer-resource model incorporates intra- and interspecific competition through the density-dependency ( $X_i$ ) of the populations and nutrients (Eq. 7-9).

I. 182: please specify "all resulting treatments" as it is still not clear what and how many mixtures were set up in addition to monocultures

A: As suggested by Reviewer 1, we changed I. 182 to:

“All the resulting 50 treatments were replicated three times for a total of 150 cultures (120 monocultures and 30 mixed cultures).” (Il. 191-192)

Exp. 1: Why did the experiment run at such a high temperature and such a low light intensity? What role could light and temperature have played for dinoflagellate performance and competition? Please comment on that.

A: We had originally planned to compare 20°C to 24°C (as a worst-case IPCC scenario for the Belgian North Sea), so experiment 1 was in fact set up in double (300 cultures). A technical issue with the climate control of the 20°C room left us without reliable data for 20°C. 24°C is, however, representative for current day summer temperatures at a local shellfish area called the “spuikom”, but this regional focus is outside the scope of article. The introduced CRM is widely applicable.

Similar to the N:P ratios, light conditions were chosen to mimic the Belgian Part of the North Sea where light is severely limited. Light penetrating the first 20 meters of the wider North Sea has a mean intensity of  $75 \mu\text{mol m}^{-2} \text{s}^{-1}$  (Gröger et al., 2013). Due to the presence of high concentrations of light-attenuating cDOM and strong mixing of the photic and euphotic zones within the Belgian EEZ, we typically expose our organisms to 20-40  $\mu\text{mol m}^{-2} \text{s}^{-1}$ . We included a reference to summer conditions in the Southern North Sea in our M&M to improve our manuscript (Il. 171-172).

Exp. 2 - please comment on the choice of nutrient concentrations and ratios (see above, from P-limitation to N-limitation, range of concentrations etc.)

A: The N:P ratios were inspired by N:P ratios occurring in the Belgian EEZ, while the L1 dilutions were inspired by the work of others, as discussed later on (II. 401-404).

Why did the authors use a different temperature (20°C as opposed to 24°C) and a different set of dinoflagellates than in Exp. 1? Only *P. micans* and *P. reticulatum* were used in both experiments, and only in the first experiment also non-HAB species were included in addition to HAB species - please clarify. Please also comment on the different experimental durations of Exp. 1 and 2

A: 20°C is more representative for summer temperatures in the Belgian EEZ. As explained above, Exp. 1 was originally designed to have both 20°C (current day) and 24°C (predicted future temp.) but we encountered technical issues with our climate room of 20°C. We used 20°C for Exp. 2 as this has more regional relevance.

Exp. 1 and Exp. 2 were performed months apart. At the time, we did not have sufficient stock of *S. trochoidea* at hand. We are aware that *P. micans* has been implicated with HABs in the past, but we considered the species to be non-HAB in accordance to the Taxonomic Reference List of Harmful Algae. The manuscript no longer refers to HAB vs. non-HAB regardless.

The duration of Exp. 1 was constrained due to practical arrangements as clarified to Reviewer 1. We had a restriction on the use of the climate room due to the problem of the other room. Exp. 2 had no such restrictions, so we could let it run longer.

II. 204/205: please clarify that a "3-species mixture" was used (see above)

A: We replaced "mixed cultures" with "3-species mixtures" as suggested (I. 213).

II. 254-256: Please explain why this simplification can be made in this context - the authors substantially vary the N:P ratios, inducing N- and P-limitation in the algal cultures, potentially influencing growth and production of secondary metabolites. Why do the authors think that N is the main driving force in determining dinoflagellate growth? They show in their first experiment that nutrient stoichiometry significantly affected the growth rate of 3 of the 4 dinoflagellates.

A: We agree that the substantial changes in the N:P ratio may have induced N- or P-limitation in our cultures. For this reason, we opted to model each N:P treatment separately rather than coming up with a single mean parameter estimates across all ratios for each species. We did not wish to suggest that nitrogen availability is the main driving force in determining dinoflagellate growth. However, we assume that growth can be adequately described by a single nutrient to predict batch culture growth. This simplification is needed to eliminate the constant of proportionality. The decision to use N-availability is strictly made due to the fact that the experiment was set up with varying nitrate concentrations. Had we modified the N:P ratio through phosphate concentrations, we could have modelled the cultures using P-availability instead. We have rephrased our M&M section to reflect that both nutrients could have been used. (II. 265-267).

I. 266: what determines the mortality coefficient?

A: The CRM's mortality coefficient is a constant used to incorporate density-dependent mortality in a simplified manner. It provides a mathematical solution that improves our model predictions in monocultures, but does not account for changes in natural mortality rates throughout the life cycle of each culture. Instead, it represents an average cell loss across the duration of the experiment. Overall, this parameter has little influence on the model predictions (estimates are very small and parameter estimates vary wildly during the convergence of the Markov chains).

## Results and Discussion:

Fig 2: please enlarge the Figure as well as axis titles and legends; do the dots represent the experimental observations? Please clarify in the figure legend.

A: We have enlarged Figure 2 and have added that markers are observations as requested. The figure can be found on page 26 of the manuscript. Fig. 1 and Fig. 3 were also reworked to enlarge their axis titles and legends.

II. 373-375: this is not true - this study was not designed to explicitly test interactions between HAB and non-HAB species, it just tested interspecific competition of a random set of dinoflagellates, some of which can form HABs

A: We removed references to competition between HAB vs. non-HAB throughout the manuscript as discussed above.

II. 404 - 405: was this also predicted from the model?

A: The CRM we used cannot predict changes in growth rate related to the availability of additional resources that are not part of its structural equations (assuming that organic nitrogen is in fact the driving mechanism here). To better highlight this issue, we have now calculated and included the coefficient of determination for *P. lima* alone (l. 360). The poor fit for *P. lima* is then highlighted in the discussion (ll. 450-454) as follows:

“CRMs can approximate the densities of both winning and losing algal species up to the plateau phase with a high degree of accuracy. Stark changes in growth rate between monocultures and multispecies cultures such as those observed *in P. lima* can, however, lead to poor predictions if the underlying mechanism is not fully understood and included in the structural equations.”

II. 415 - 421: The authors should state that already in the introduction to clarify the motivation of their study

A: The abstract and introduction were changed to put more emphasis on the model. The N:P ratio is now only mentioned once at the end of the introduction to make sure that readers do not get false expectations.

I. 421 - 422: see comment above - please explain why this simplification is suitable here.

A: As discussed above, we rephrased the M&M section (ll. 265-267) to reflect that growth can be described by either nutrient and that we made a choice to use nitrogen. We then further highlight that both are optional in the discussion. The paragraph (ll. 445-447) now reads:

“In this study, the N:P ratios and the CF's were merely used to introduce variability in the nitrate concentrations, which we then chose as the driver of the consumer-resource model used.”

I. 424-425: see above - in the context of this study it is completely irrelevant whether the dinoflagellates are HAB or non-HAB species as this is neither considered in the experiments nor in the model. Statements like this mislead the reader to think that this was explicitly tested!

A: We agree with the reviewer and have removed all references to HAB vs. non-HAB throughout the manuscript when discussing our study.

I. 429 - 430: It is not possible for the authors to tell whether direct interactions played a role at all, because none of them were tested. The species / strain specific traits are not known (or have not been introduced)

A: We agree that the previous statement was too bold and have removed the reference to direct interactions from the sentence. The corresponding section now states:

“By using a CRM, this study was able to demonstrate that the presence of a fast-growing species (*P. micans*) had strong, indirect negative effects on the growth of competing dinoflagellates; the growth of competing algae was to a large degree hampered by diminishing nutrient availability due to uptake by *P. micans*.” (ll. 454-457).

II. 431 - 433: This should be stated in the introduction; however, there is a high intraspecific variability in the production of allelochemicals in many dinoflagellates and the fact that all species have been found to potentially produce allelochemicals does not mean that the strains used here were able to do so

A: Considering all comments, we decided not to include an overview of the allelopathic properties of each species in the introduction to avoid creating false expectations from the reader. As the reviewer rightfully points out, we did not explicitly test for allelopathy in the current study. We do, however, believe that we need to discuss allelopathy in the introduction and discussion to highlight the strengths and weaknesses of the CRM, and to provide recommendations for the next iteration of CRMs. To be more clear in the manuscript, we now state that we did not explicitly test the ability of our strains to produce allelochemicals (l. 460).

II. 433 - 442: this is possible, but cannot be deduced from the present study - see above

A: We have removed the text related to our earlier attempts at modelling our observations using Lotka-Volterra based approaches as, indeed, these faulty analyses were not shown here.

II. 475 - 477: this should not be stated at the very end of the discussion, but in the introduction to make clear what this study aims at and potentially can provide and what not (now II. 498-500)

A: As discussed above, the manuscript now puts more emphasis on CRMs being a starting point for future trait-based modelling. By working with our basic CRM on this dataset, we conceptually developed an extended model that includes allelopathy and might be relevant for future research. This theoretical framework did not exist at the time that we designed our experiments, as it is part of the lessons learned during this study, so our data was never intended to suit the model that we propose in the discussion. It is, however, likely that other researchers may come up with different ways of extending the basic CRM in ways that avoid the need for bi-algal cultures.

## Highlights

- Five common dinoflagellates were co-cultured under 22 nutrient regimes.
- Monoculture growth was used to parametrize a consumer-resource model (CRM).
- Consumer-resource modelling can predict species dominance in mixed batch cultures.
- CRMs may differentiate resource assimilation, resource efficiency, and natural mortality.

## Abstract

Global change will disturb the frequency, scale and distribution of harmful algal blooms (HABs), but we are unable to predict future HABs due to our limited understanding of how physicochemical changes in the environment affect interspecific competition between dinoflagellates. Trait-based mechanistic modelling is an important tool to unravel and quantify various direct and indirect interactions between species. The present study explores whether MacArthur's consumer-resource model can be used as a viable base model to predict dinoflagellate growth in closed multispecies systems. To this end, two batch culture experiments (294 cultures in total) with monocultures and multispecies cultures of *Alexandrium minutum*, *Prorocentrum lima*, *P. micans*, *Protoceratium reticulatum* and *Scrippsiella trochoidea* were performed. Despite changes to the relative (different nitrate concentrations) and absolute nutrient availability (dilutions of L1 medium), *P. micans* outcompeted all other species in mixed cultures. Consumer-resource modelling parameterized using monoculture growth correctly predicted this species dominance ( $R^2$  between 0.80 and 0.95). Parameter estimates revealed that *P. micans* had a faster uptake of nitrogen when compared to its competitors, but did not differ in resource efficiency and natural mortality rate. Yet, while the model accurately predicted community dynamics during the growth phase, it was not able to predict their dynamics beyond the point of quiescence. Consumer-resource modelling was shown to differentiate the roles of resource assimilation, resource efficiency, and natural mortality rates in batch culture experiments with minimal data requirements beyond common measurements. The results suggest that consumer-resource models provide a promising basis for trait-based modelling of interspecific competition between (harmful) algae.

1 **Title: Monoculture-based consumer-resource models predict species dominance in mixed**  
2 **batch cultures of dinoflagellates.**

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22

23 Declarations of interest: none

24

25 Running title:

26 Nutrient competition in mixed batch cultures of dinoflagellates: species dominance predicted by  
27 a simple consumer-resource model

28

29 **Keywords:** Consumer-resource modelling, mixed batch cultures, dinoflagellates

## 30 **Abstract**

31 Global change will disturb the frequency, scale and distribution of harmful algal blooms (HABs),  
32 but we are unable to predict future HABs due to our limited understanding of how physicochemical  
33 changes in the environment affect interspecific competition between dinoflagellates. Trait-based  
34 mechanistic modelling is an important tool to unravel and quantify various direct and indirect  
35 interactions between species. The present study explores whether MacArthur's consumer-  
36 resource model can be used as a viable base model to predict dinoflagellate growth in closed  
37 multispecies systems. To this end, two batch culture experiments (294 cultures in total) with  
38 monocultures and multispecies cultures of *Alexandrium minutum*, *Prorocentrum lima*, *P. micans*,  
39 *Protoceratium reticulatum* and *Scrippsiella trochoidea* were performed. Despite changes to the  
40 relative (different nitrate concentrations) and absolute nutrient availability (dilutions of L1  
41 medium), *P. micans* outcompeted all other species in mixed cultures. Consumer-resource  
42 modelling parameterized using monoculture growth correctly predicted this species dominance  
43 ( $R^2$  between 0.80 and 0.95). Parameter estimates revealed that *P. micans* had a faster uptake of  
44 nitrogen when compared to its competitors, but did not differ in resource efficiency and natural  
45 mortality rate. Yet, while the model accurately predicted community dynamics during the growth  
46 phase, it was not able to predict their dynamics beyond the point of quiescence. Consumer-  
47 resource modelling was shown to differentiate the roles of resource assimilation, resource  
48 efficiency, and natural mortality rates in batch culture experiments with minimal data requirements  
49 beyond common measurements. The results suggest that consumer-resource models provide a  
50 promising basis for trait-based modelling of interspecific competition between (harmful) algae.

## 51 **Highlights**

- 52 • Five common dinoflagellates were co-cultured under 22 nutrient regimes.
- 53 • Monoculture growth was used to parametrize a consumer-resource model (CRM).
- 54 • Consumer-resource modelling can predict species dominance in mixed batch cultures.
- 55 • CRMs may differentiate resource assimilation, resource efficiency, and natural mortality.

## 56 **1. Introduction**

57 Phycologists have tried to understand and predict the spatiotemporal occurrence of harmful algal  
58 blooms (HABs) for decades. Red tides were first considered to be inherently unpredictable due  
59 to the dynamic nature of marine ecosystems as well as the vast number of functional properties  
60 (e.g. nutrient uptake rates, internal storage, pigment composition etc.) and adaptive strategies  
61 (e.g. cyst production, cell shape, motility, thin layer formation) the causative organisms may have  
62 (Sweeney, 1978, 1975). Over the years, it was discovered that phytoplankton communities are  
63 structured by nutrient competition, species interactions (grazing, allelopathy), abiotic variables  
64 (light, temperature, turbulence etc.) and stochastic processes (Armstrong, 1979; Eppley, 1972;  
65 Huisman and Weissing, 1994; Legrand et al., 2003; Margalef, 1978; Richerson et al., 1970;  
66 Smayda, 2008; Tilman, 1977). Today, it is widely accepted that HAB development results from  
67 exceptional successions of phytoplankton that require specific environmental conditions to occur  
68 (Stoecker et al., 2008). Identifying the sets of biotic and abiotic conditions that enable the initiation  
69 and development of HABs, sometimes referred to as “windows of opportunity”, has been a major  
70 goal of HAB research from the start.

71 Ramón Margalef observed that nutrient availability and the decay of turbulent energy determine  
72 the succession of groups of phytoplankton and, hence, the likelihood of toxic bloom development  
73 (Margalef, 1978). In his now-famous “mandala”, harmful red tides may develop when the nutrient  
74 availability is high and the turbulent energy is restricted. While his mandala was improved through  
75 the addition of functional properties, demographic strategies and the inclusion of novel HAB taxa  
76 (e.g. Allen and Polimene, 2011; Balch, 2004; Glibert, 2016), neither the original mandala nor the  
77 recent renditions were able to resolve the non-deterministic nature of HAB development. Blooms  
78 often fail to develop under seemingly ideal conditions. To this day, we are unable to reliably predict  
79 how changes in either the relative or absolute availability of nutrients affect the risk of HABs in a  
80 given phytoplankton community. While both chronic and episodic eutrophication seem to be linked  
81 to HABs, there is no clear evidence that nutrients promote HABs by themselves. Nutrient uptake

82 kinetics and resource preferences vary greatly within and between HAB species (Glibert and  
83 Burkholder, 2006) and cannot be distinguished from those of closely related non-HAB species  
84 (Anderson et al., 2002; Heisler et al., 2008; Wells et al., 2015). This hampers our ability to predict  
85 where and when HABs will occur based on resource abundance alone. It is, however, clear that  
86 eutrophication affects the entire food web, altering every biological interaction (e.g. nutrient  
87 competition, grazing, allelopathy) that collectively determines the success of harmful algae  
88 (Glibert et al., 2010; Granéli et al., 2008b; Smayda, 2008).

89 Dinoflagellates are poor competitors for nutrients and, hence, are at risk of competitive exclusion  
90 (Smayda, 1997). They also face strong grazing control by microzooplankton, mesozooplankton  
91 and benthic filter feeders (Smayda, 2008; Tillmann, 2004; Turner, 2006). In order to cope with  
92 both these interspecific interactions, dinoflagellates have evolutionary adaptations such as the  
93 production of cysts, mixotrophy, (toxin-mediated) allelopathy and grazer deterrence (Bravo and  
94 Figueroa, 2014; Chakraborty et al., 2015; Crane and Grover, 2010; Roy and Chattopadhyay,  
95 2007). Toxins, grazer deterrents and allelochemicals - i.e. exudates that cause nutrient leakage,  
96 inhibit photosynthesis, arrest the cell-cycle, or affect other enzymes of competing algae (Granéli  
97 and Hansen, 2006; Legrand et al., 2003; Reigosa et al., 1999) - reduce the long-term extinction  
98 risk of toxic algae and may help maintain toxic blooms (Granéli et al., 2008a; Ianora et al., 2011;  
99 Smayda, 1997; Smayda, 2008; Turner, 2006; Xu and Kiørboe, 2018). Allelopathy and grazer  
100 deterrence should allow increasingly dominant organisms to overpower their competitors during  
101 HAB initiation. Yet, to date, their role during the first stages of HAB development remains unclear.

102 Toxic effects are variable or inducible in nature (e.g. Dam and Haley, 2011; Poulin et al., 2018),  
103 mostly occur at bloom-level densities (Jonsson et al., 2009), and can have a high individual cost  
104 for toxic species while providing collective benefits to others (Driscoll et al., 2016; Flynn, 2008a).  
105 For these reasons, there is doubt that these chemical interactions play a crucial role during bloom  
106 initiation. More recently, Blossom et al. (2019) have demonstrated that allelochemicals can yield  
107 significant cell-level benefits at very low cell concentrations. Moreover, they also suggested that

108 meaningful trade-offs between allelopathy and growth rate (i.e. fitness costs) determine whether  
109 allelochemicals are released. Overall, these studies demonstrate that the processes behind  
110 allelopathy need to be unravelled further to understand the non-deterministic nature of HABs  
111 during windows of opportunity.

112 Allelopathic interactions between microalgae are usually studied in one of three ways: (1) through  
113 the addition of cell-free culture filtrates to competitors; (2) by using caged batch cultures whereby  
114 both species are co-cultured, but separated by a permeable mesh or membrane; (3) by means of  
115 co-existence experiments that co-culture both species in direct contact. Each method has its own  
116 drawbacks. The toxicity of intact cells can sometimes be increased by or be dependent on direct  
117 contact with targets (Driscoll et al., 2016). As a result, caution should be used when interpreting  
118 the results of the first two methods. Co-existence experiments, on the other hand, do not separate  
119 chemical interactions (i.e. allelopathy) from other interactions such as resource competition and  
120 mixotrophy (Allen et al., 2016). This issue is often addressed by use of the dilution method. By  
121 increasing the number of target cells relative to a constant density of an allelopathic species, the  
122 amount of allelochemicals per target cell decreases. This should lead to increased growth of the  
123 target species. If the growth rate remains constant or decreases due to an increase in competition,  
124 allelopathy is considered to be absent (Weidenhamer, 2006). Crucially, this approach fails to  
125 address two key aspects of allelopathic interactions: (1) that they may vary from strongly inhibitory  
126 to negligible to stimulatory between strains (Poulin et al., 2018), and (2) that they can be induced  
127 by increased nutrient competition (Granéli et al., 2008b), obscuring the interpretation of the test.  
128 Mechanistic modelling of culture dynamics may help alleviate these problems and could improve  
129 our understanding of interspecific interactions.

130 The first mathematical description of allelopathy in phytoplankton, where an interaction term was  
131 added to a two species Lotka-Volterra model, was proposed by Maynard Smith (1974). Over the  
132 years, numerous improvements and refinements were made to the Maynard-Smith function (e.g.  
133 Bandyopadhyay, 2006; Chattopadhyay, 1996; Mandal et al., 2014; Mukhopadhyay et al., 2003,

134 1998; Solé et al., 2005), demonstrating the high potential of Lotka-Volterra derived models. Yet,  
135 despite their merits, the Maynard-Smith-based models are completely dependent on direct,  
136 density-dependent interactions to assess intraspecific and interspecific competition. Neither the  
137 Lotka-Volterra equations, nor the Maynard-Smith equations, include spatiotemporal dynamics of  
138 nutrients. Considering that resource competition is a major determinant of blooms (Sourisseau et  
139 al., 2017), this study explores whether models that describe how consumers interact indirectly  
140 through the use of common resources can be used as an alternative approach.

141 MacArthur and Levins (1967) introduced resource utilization functions into the Lotka-Volterra's  
142 equations, which was later developed into a consumer-resource model (MacArthur, 1970, 1969).  
143 In contrast to Maynard-Smith's later model, MacArthur's consumer-resource model (CRM) does  
144 not include density-dependent species interactions. Instead, species interact exclusively by using  
145 shared resources (section 2.5). This model shares some commonality with the Rosenzweig-  
146 MacArthur consumer-resource model, but strives towards a simplification of resource competition  
147 dynamics. While it was quickly rejected as a suitable method for understanding niche overlap  
148 within natural environments (Abrams, 1975), the model garnered attention as a sound basis for  
149 theoretical work (Chesson, 1990). Consumer-resource models have since been used to describe  
150 competition dynamics in various organisms.

151 This study investigates whether consumer-resource models, like Maynard-Smith based models,  
152 may function as valuable base models to unravel competition among co-occurring dinoflagellates.  
153 Five dinoflagellates that are found in the North Sea (*Alexandrium minutum*, *Prorocentrum lima*,  
154 *P. micans*, *Protoceratium reticulatum* and *Scrippsiella trochoidea*) were grown in 294 single and  
155 multispecies cultures spread across two experiments. Various nutrient treatments (varying either  
156 the N:P ratio, the order of magnitude of nutrient concentrations, or both) were used to determine  
157 whether a CRM can reproduce resource competition in multispecies cultures of dinoflagellates  
158 under different nutrient regimes. The initial growth and species dominance were then shown to  
159 be largely predictable using consumer-resource modelling based on monoculture behaviour.

160

## 161 2. Material and Methods

### 162 2.1 Stock cultures

163 *Alexandrium minutum* (SCCAP K-0993) and *Protoceratium reticulatum* (SCCAP K-1478) were  
164 bought from the Scandinavian Culture Collection of Algae & Protozoa (Copenhagen, Denmark).  
165 *Prorocentrum lima* (CCAP1136/9) and *P. micans* (CCAP1136/20) were obtained from the Culture  
166 Collection of Algae and Protozoa (Oban, Scotland). *Scrippsiella trochoidea* is an in-house strain,  
167 isolated from the Belgian Part of the North Sea one year before the experiments. Stock cultures  
168 of all dinoflagellates were grown in L1 medium, prepared from Instant Ocean™ artificial seawater  
169 (Belcopet, Belgium) in accordance with Guillard and Hargraves (1993) and replenished ( $\pm$  80%)  
170 every 2 weeks. Cultures were grown at 20°C with a 12-hour light-dark cycle ( $20\text{-}40 \mu\text{mol}\cdot\text{m}^{-2}\cdot\text{s}^{-1}$ ),  
171 similar to summer conditions in the photic zone of the Southern North Sea (Gröger et al., 2013;  
172 Mortelmans et al., 2019). The growth of stock cultures was monitored through weekly cell counts  
173 using a Sedgewick-Rafter counting chamber and a Kyowa Optical Biolux-2 light microscope. Both  
174 experiments used cells taken from stock cultures that were growing exponentially.

175

### 176 2.2 Experiment 1: 4 species, 10 N:P ratios

177 A first experiment was set up to investigate whether small variations in nutrient availability and  
178 nutrient stoichiometry affect the interspecific competition among dinoflagellates in batch cultures.  
179 Ten algal growth media were prepared so to have ten unique nitrogen-to-phosphorus ratios.  
180 Regular L1 medium contains  $882 \mu\text{M NO}_3^-$  and  $36.2 \mu\text{M PO}_4^{3-}$ , corresponding to a N:P ratio of 24.  
181 By only adjusting the  $\text{NO}_3^-$  concentrations, ten growth media with different nitrogen-to-phosphorus  
182 ratios were prepared: preparations of 294, 368, 441, 478, 515, 551, 588, 662, 735 or  $882 \mu\text{M NO}_3^-$   
183 corresponded to a N:P of 8, 10, 12, 13, 14, 15, 16, 18, 20 or 24, respectively. All other components  
184 of L1 medium ( $\text{PO}_4^{3-}$ , trace metals, vitamins) were added at the regular dose. The range of N:P  
185 ratios that was used was based on N:P ratios that were observed in the Belgian part of the North

186 Sea between 2013 and 2019 (Mortelmans et al., 2019). On average, the Belgian EEZ has a mean  
187 N:P ratio of 22 while the average N:P ratio in a local shellfish area is around 14 during summer.  
188 Monocultures of *P. micans*, *P. lima*, *S. trochoidea*, and *P. reticulatum* were set up in each media  
189 by adding 100 cells ml<sup>-1</sup> to Erlenmeyer flasks filled with 50 ml of medium. Mixed cultures were set  
190 up in each medium by adding 100 cells ml<sup>-1</sup> of each of the four algae to 50 ml of medium. All the  
191 resulting 50 treatments were replicated three times for a total of 150 cultures (120 monocultures  
192 and 30 mixed cultures).

193 Cells were grown for 28 days at 24°C with a 12-hour photoperiod of 30±5 μmol m<sup>-2</sup> s<sup>-1</sup>. Twice a  
194 week, a 1 ml was taken from each flask, fixed with 100 μl of 12% formaldehyde, and counted  
195 using a Sedgewick-Rafter counting chamber and light microscopy. Additional samples (7 ml) were  
196 taken on day 14 and day 28 for nutrient analyses. During the first experiment, the NO<sub>3</sub> and PO<sub>4</sub>  
197 concentrations were determined using spectrophotometric test kits (Merck Millipore, Darmstadt,  
198 Germany) that need large volumes (20 ml). For this reason, replicates were pooled and filtered  
199 with Millex-GV 0.22 μm PVDF syringe filters (Merck Millipore). Filtrates and initial media (day 1)  
200 were then analysed using an Aquamate spectrophotometer (Thermo Scientific, San Jose, USA).

201

### 202 2.3 Experiment 2: 3 species, 3 N:P ratios, 4 orders of magnitude

203 A second experiment was performed to examine whether the interspecific competition between  
204 dinoflagellates in batch cultures is affected by larger differences in macronutrient availability.  
205 Twelve unique algal growth media were made based on regular L1 growth medium. Media were  
206 first prepared with 294, 588 or 882 μM NO<sub>3</sub><sup>-</sup> to obtain three N:P ratios (8, 16 and 24). All other L1  
207 components (PO<sub>4</sub><sup>3-</sup>, vitamins and trace elements) were added at the regular dose. Each medium  
208 was subsequently diluted by a factor 1, 10, 100 or 1000 to obtain media with 100%, 10%, 1% or  
209 0.1% volume fractions of L1 medium vs. Instant Ocean™ artificial seawater. Hereafter, these  
210 dilutions will be referred to as “concentration factors” (CFs), so that the medium with a N:P ratio  
211 of 24 and a CF of 100 reflects actual L1 medium, while the medium with a N:P ratio of 8 and a CF

212 of 0.1 corresponds to 0.033% of L1 medium. In each of the twelve resulting media, monocultures  
213 of *A. minutum*, *P. reticulatum* and *P. micans*, as well as 3-species mixtures of these algae, were  
214 made by adding 100 cells.mL<sup>-1</sup> (each) to 75 ml of medium in Erlenmeyer flasks. Every treatment  
215 was replicated three times, resulting in 144 cultures (108 monocultures and 36 mixed cultures).  
216 Cultures were placed at 20°C with a 12-hour photoperiod of 33±6 μmol.m<sup>-2</sup>.s<sup>-1</sup> for 56 days. Twice  
217 a week, cell counts were made using a Sedgewick-Rafter counting chamber and light microscopy.  
218 Once a week, 2 ml samples of each flask were taken for nutrient analysis. For this experiment,  
219 we used a QuAAtro segmented flow analyser to determine the N-NO<sub>3</sub> and P-PO<sub>4</sub> concentrations  
220 using the colorimetric methods found in Hansen and Koroleff (1999). Around 5 ml was needed for  
221 both analyses. To this end, replicates were filtered and pooled as described for experiment 1. The  
222 filtrates were stored at 4°C in 15 ml falcon tubes prior to their analysis.

223

#### 224 2.4 Simple growth models

225 Growth rates ( $\mu$ ; d<sup>-1</sup>) and carrying capacities ( $K$ ; μm<sup>3</sup>.ml<sup>-1</sup>) were determined to assess the overall  
226 growth of each dinoflagellate. Cell counts ( $N_t$ ) were transformed to biovolume (μm<sup>3</sup>.ml<sup>-1</sup>) using  
227 size measurements and the geometric formulas of Olenina et al. (2006). The conversion factors  
228 (μm<sup>3</sup>.cell<sup>-1</sup>) used were: 7299 (*A. minutum*), 7580 (*S. trochoidea*), 12596 (*P. reticulatum*), 20293  
229 (*P. micans*), and 43960 (*P. lima*). Depending on whether or not the stationary growth phase was  
230 reached, the biovolumes of each flask were fitted with exponential or logistic growth models using  
231 least square optimisation in the 'nls' function in R (Baty et al., 2015). Multiple group comparisons  
232 by means of Kruskal Wallis (KW) tests (Kruskal and Wallis, 1952) were used to compare growth  
233 parameters ( $\mu$  and  $K$ ) between treatments (N:P) and species. Pairwise comparisons using Dunn's  
234 multiple comparison (DMC) test (Dunn, 1964) were made to investigate the effects of treatments  
235 (CF, mono vs. mixed) on the growth of each species. Linear regression models (LM) were used  
236 to detect linear responses to nutrient stoichiometry as described by Wilkinson and Rogers (1973).

237

238 2.5 Development of a community model

239 An adaptation of MacArthur's (1970) consumer-resource model for non-interacting resources was  
 240 used to predict competition between dinoflagellates in mixed batch cultures using only the uptake  
 241 and conversion of nutrients by individual species. According to the original model, predators ( $n$ )  
 242 interact solely by consuming common, non-interacting prey species ( $k$ ). As a result, the per capita  
 243 growth rate of a predator ( $i$ ) can be described by the following equation (Eq. 1):

$$244 \quad (1) \frac{1}{X_i} \cdot \frac{dX_i}{dt} = C_i \cdot \left( \sum_{k=1}^n a_{i,k} \cdot w_k \cdot R_k - T_i \right)$$

245 Where  $X_i$  is the population density of the predator  $i$ ;  $R_k$  is the population density of prey species  $k$ ;  
 246  $a_{i,k}$  is the probability that predator  $i$  captures prey species  $k$ ;  $w_k$  is the weight of prey species  $k$ ;  
 247 and  $T_i$  is the threshold weight that the predator needs to capture per capita to get a net population  
 248 growth of 0 (MacArthur, 1970). Any excess of prey captured (i.e. the result of the sum) is  
 249 converted to population growth by a constant of proportionality  $C_i$  that governs the conversion of  
 250 grams of resource captured to grams of  $X_i$ . Because of predation, the logistic population growth  
 251 of prey species  $k$  is reduced by consumer-imposed mortality (Eq. 2), with  $r_k$  being the growth rate  
 252 of prey species  $k$  and  $K_k$  being the carrying capacity of its environment.

$$253 \quad (2) \frac{1}{R_k} \cdot \frac{dR_k}{dt} = r_k \cdot \left( 1 - \frac{R_k}{K_k} \right) - \left( \sum_{i=1}^n a_{i,k} \cdot X_i \right)$$

254 Here, we propose that MacArthur's consumer-resource model can be adapted to the uptake of  
 255 non-interacting abiotic nutrients by describing resource abundance as:

$$256 \quad (3) \frac{1}{R_k} \cdot \frac{dR_k}{dt} = I_k - \left( \sum_{i=1}^n a_{i,k} \cdot X_i \right)$$

257 Where  $I_k$  is the renewal of resources by riverine discharge, submarine weathering, atmospheric  
 258 exchange, and biological activity (remineralisation, nitrogen fixation etc.). In closed environments  
 259 like our batch cultures the short-term renewal of resources was assumed to be negligible ( $I_k = 0$ ).

260 When applied to the present setup, i.e. dinoflagellates interacting through the consumption of  
 261 nitrate and phosphate, the following equations were derived from the model:

$$262 \quad (4) \frac{1}{X_i} \cdot \frac{dX_i}{dt} = C_i \cdot (a_{i,NO_3} \cdot w_{NO_3} \cdot [NO_3^-] + a_{i,PO_4} \cdot w_{PO_4} \cdot [PO_4^-] - m_i)$$

$$263 \quad (5) \frac{1}{[NO_3^-]} \cdot \frac{d[NO_3^-]}{dt} = - \sum_{i=1}^n a_{i,NO_3} \cdot X_i$$

$$264 \quad (6) \frac{1}{[PO_4^-]} \cdot \frac{d[PO_4^-]}{dt} = - \sum_{i=1}^n a_{i,PO_4} \cdot X_i$$

265 The model was simplified to a prototypical consumer-resource model by assuming that growth  
 266 can be adequately described by either nutrient. Here, nitrogen was used (i.e.  $w_{PO_4}$  was assumed  
 267 to be 0) since the experimental design included most variability in nitrogen concentrations. Next,  
 268 the constant of proportionality  $C_i$  was merged with the parameters  $w_{NO_3}$  and  $m_i$ . In the end, the  
 269 uptake and conversion of nitrogen was used to predict the growth of each dinoflagellate (Eq. 7,  
 270 Eq. 8). Phosphorus measurements were used to estimate the uptake of phosphorus (Eq. 9) using  
 271 the predicted per capita growth. The final model is:

$$272 \quad (7) \frac{dX_i}{dt} = X_i \cdot (U_{i,NO_3} \cdot W_{NO_3} \cdot [NO_3^-] - M_i)$$

$$273 \quad (8) \frac{d[NO_3^-]}{dt} = -[NO_3^-] \cdot \sum_{i=1}^n U_{i,NO_3} \cdot X_i$$

$$274 \quad (9) \frac{d[PO_4^-]}{dt} = -[PO_4^-] \cdot \sum_{i=1}^n U_{i,PO_4} \cdot X_i$$

275 Where  $X_i$  is the density (in biovolume) of dinoflagellate  $i$  ( $\mu\text{m}^3 \cdot \text{l}^{-1}$ );  $U_{i,NO_3}$  is the probability of uptake  
 276 of  $\text{NO}_3$  per dinoflagellate  $i$  per time unit ( $\text{d}^{-1}$ );  $W_{NO_3}$  is the conversion efficiency, i.e. the biovolume  
 277 formed by dinoflagellate  $i$  per unit  $\text{NO}_3$  taken up ( $\mu\text{m}^3 \cdot \mu\text{g}^{-1}$ );  $M_i$  is a mortality coefficient (i.e. the  
 278 fraction of biovolume dinoflagellate  $i$  loses daily;  $\mu\text{m}^3 \cdot \text{d}^{-1}$ );  $[NO_3^-]$  is the abundance of  $\text{NO}_3$  ( $\mu\text{g}$ );  
 279  $C_{PO_4}$  is the abundance of  $\text{PO}_4$  ( $\mu\text{g}$ );  $U_{i,PO_4}$  is the probability of uptake of  $\text{PO}_4$  per unit of  
 280 dinoflagellate  $i$  per time unit ( $\text{d}^{-1}$ ).

281

## 282 2.6 Applying the model

283 Monoculture data was used to estimate the parameters ( $U_{NO_3}$ ,  $U_{PO_4}$ ,  $W_{NO_3}$ ,  $M_i$ ) per treatment and  
284 dinoflagellate with a simulated annealing algorithm. The mean absolute percentage error was  
285 used as an objective function to ensure an equal fit across the different magnitudes of species'  
286 densities. Markov chain Monte Carlo (MCMC) simulations (Hastings, 1970) were then used to  
287 generate the joint posterior distributions for each parameter. The parameter space was restricted  
288 to 50% deviation of the initial estimates to get fast parameter convergence. Convergence of the  
289 posterior distributions of three parallel Markov chains was assessed based on the Gelman-Rubin  
290 convergence criterion (Gelman and Rubin, 1992) and plotted to manually optimize burn-in.  
291 Predictions for both monocultures and mixtures cultures (densities and nutrients) were obtained  
292 using a 1000 Monte Carlo simulations, each randomly drawing parameter estimates from the  
293 posterior distributions. Predictions of each simulation were stored. Model performance was  
294 assessed by comparing the observed species densities to the median predicted densities. All  
295 calculations were done in the statistical software R using the deSolve (Soetaert et al., 2010),  
296 abind (Plate and Heiberger, 2011), and GenSA (Xiang et al., 2013) packages.

297

298 **3. Results**

## 299 3.1. Relative resource availability

300 During the first experiment all cultures were grown for 28 days. Logistic growth models were used  
301 to determine the monoculture growth rates for all species except *P. lima*, which was still growing  
302 exponentially at the end of the experiment. Exponential growth models were used to determine  
303 the growth rates of *P. lima* instead (Supporting figures SF1-4). Overall, *P. micans* had the highest  
304 growth rate ( $0.46 \pm 0.07 \text{ d}^{-1}$ ;  $\mu \pm \sigma$ ), followed by *S. trochoidea* ( $0.37 \pm 0.04 \text{ d}^{-1}$ ), *P. reticulatum*  
305 ( $0.28 \pm 0.04 \text{ d}^{-1}$ ), and *P. lima* ( $0.04 \pm 0.01 \text{ d}^{-1}$ ). Nutrient stoichiometry significantly affected the growth  
306 rate of *P. micans*, *P. reticulatum* and *S. trochoidea* (KW  $P < 0.05$ ), but not *P. lima* ( $P > 0.05$ ), in a

307 nonlinear fashion (Fig. 1A). A significant linear relationship (LM  $P < 0.001$ ) was found between  
308 the initial N:P ratio and the carrying capacities of *P. micans*, but not between the N:P ratio and  
309 the carrying capacities of *P. reticulatum* or *S. trochoidea* ( $P > 0.05$ ). On average, the carrying  
310 capacity of *P. reticulatum* ( $8.6 \pm 3.9 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$ ) was significantly higher than those of *P. micans*  
311 ( $5.5 \pm 1.4 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$ ) and *S. trochoidea* ( $4.4 \pm 1.1 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$ ; Fig 1B). *P. micans* outcompeted  
312 all other species in all mixed cultures (supporting figure SF5) while maintaining growth rates which  
313 were similar (DMC  $P > 0.05$ ) to those in monoculture ( $0.43 \pm 0.03 \text{ d}^{-1}$ ;  $\mu \pm \sigma$ ).  
314 No significant effect of the N:P ratio on the growth rate of *P. micans* in mixed cultures was found  
315 (KW  $P > 0.05$ ), but the linear effect of the N:P ratio on the carrying capacity of *P. micans* persisted.  
316 On average, *P. micans* lost over half of its carrying capacity to competitors: its average carrying  
317 capacity decreased from  $5.5 \pm 1.4 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$  in monocultures to  $2.1 \pm 0.4 \cdot 10^8 \mu\text{m}^3 \cdot \text{ml}^{-1}$  in mixed  
318 cultures. *S. trochoidea* and *P. reticulatum* both reached peak density around day 14, after which  
319 densities plateaued or declined. Exponential or logistic growth models were used to determine  
320 their initial growth rates (up to day 17). These were  $0.44 \pm 0.10$  and  $0.31 \pm 0.15 \text{ (d}^{-1}\text{)}$  for *S.*  
321 *trochoidea* and *P. reticulatum*, respectively. Neither were statistically different from monoculture  
322 growth rates (DMC  $P > 0.05$ ). Uniquely, *P. lima* grew faster in mixed cultures; it grew at a growth  
323 rate of  $0.09 \pm 0.01 \text{ d}^{-1}$  for the duration of the experiment. Nitrogen and phosphorus concentrations  
324 from the first experiment can be found in supporting figures SF6-10. In mixed cultures, nutrients  
325 were depleted in all but the highest N:P ratio by day 14 (Fig. SF10).

326

### 327 3.2 Absolute (and relative) resource availability

328 The second experiment lasted 56 days, but the lowest concentration factors (CF0.1 and CF1) did  
329 not support prolonged growth. We observed between 1 (the lowest belonging to *P. reticulatum*)  
330 and 3 (found for *A. minutum*) population doublings before growth stalled. These treatments were  
331 no longer sampled after 39 days. Logistic growth models were used (Supporting Figure SF11) to  
332 determine the growth rate and carrying capacity (Table 1) of CF10 and CF100 monocultures. The

333 mean growth rate of *P. micans* ( $0.31 \pm 0.04 \text{ d}^{-1}$ ;  $\mu \pm \sigma$ ) exceeded the growth rates of *A. minutum*  
334 ( $0.27 \pm 0.03 \text{ d}^{-1}$ ) and *P. reticulatum* ( $0.19 \pm 0.03 \text{ d}^{-1}$ ). Growth rates were usually higher at CF10  
335 (KW  $P < 0.001$ ), and did not differ between N:P ratios (LM  $P > 0.05$ ). The N:P ratio did have a  
336 significant (LM  $P < 0.01$ ) positive effect on the carrying capacities of the three dinoflagellates at  
337 both CF10 and CF100.

338 *P. micans* dominated all multispecies cultures (Supporting figure SF12). The growth rates of each  
339 species were determined by logistic growth models for CF10 and CF100. No significant  
340 differences were found between the growth rates of monocultures and mixed cultures for any of  
341 the three species (KW  $P > 0.05$ ). The N:P ratio had no effect on the growth rate of any of the  
342 dinoflagellates at neither CF (KW  $P > 0.05$ ; Table 1), but the linear effect of the N:P ratio on the  
343 carrying capacity of *P. micans* was again found (LM  $P < 0.05$ ). Nutrients were depleted between  
344 day 20 and day 25 in virtually all cultures (Supporting figure SF16).

345

### 346 3.3 Consumer-resource modelling

347 Overall, our consumer-resource model was able to predict most of the variation in abundance of  
348 monocultures of both experiments; the coefficients of determination ( $R^2$ ) were 0.8981 and 0.9765  
349 for monoculture growth in the first and second experiment, respectively (Fig. 3). During the first  
350 experiment, *P. micans* and *S. trochoidea* – the two species that grew fastest and, hence, most in  
351 mixed cultures – were found to have similar nitrogen conversion efficiencies and likelihoods of  
352 nitrogen uptake (Table 2). By contrast, *P. reticulatum* exhibited a markedly lower likelihood of  
353 nitrogen uptake and a higher nitrogen conversion efficiency. *P. lima* had a nitrogen conversion  
354 efficiency that far exceeded those of all other species, which could be a computational artefact.  
355 When used to predict abundances for the entire duration of the first experiment, the goodness-of-  
356 fit of the model was generally poor ( $R^2 = 0.3581$ ). Yet, when looking at the data up to quiescence,  
357 which we isolated by first identifying the highest density per species and then removing all counts  
358 after  $t_{\max}$  which were smaller than 80% of the peak abundances, we found that the model generally

359 produced good predictions for the exponential growth of mixed cultures ( $R^2 = 0.8191$ ; all species).  
360 Densities of *P. lima* in mixed cultures were, however, predicted poorly ( $R^2 = 0.12$ ).  
361 The increased temporal resolution of nutrient data during the second experiment greatly improved  
362 the model's performance in mixed cultures. When used to predict the growth of mixed cultures of  
363 the two highest concentrations factors (CF10 and CF100), a coefficient of determination of 0.8910  
364 was found for all data. Using the same quiescence filter as before to remove the death phase, the  
365 goodness-of-fit improved even further ( $R^2 = 0.9289$ ; Fig. 2-3). Overall, the natural mortality rates  
366 during exponential growth were negligible for all dinoflagellates and did not differ greatly between  
367 species and experiments. In addition, we generally found that changes in growth rates – such as  
368 those linked to CF's and NP ratios – are coupled to differences in the likelihood of nutrient uptake  
369 (LM:  $P < 0.001$  for exp. 1; LM  $P < 0.01$  for exp. 2), but not to changes in the nitrogen conversion  
370 efficiencies ( $P \geq 0.05$  for both experiments).

371

#### 372 **4. Discussion**

373 Despite decades of experimental and observational research, much can still be learned of the key  
374 biological processes that influence HAB development. Interspecific competition between (closely)  
375 related species, in particular, is far from fully understood (Wells et al., 2015). Even though several  
376 key biological processes are known to affect HAB development (e.g. grazer resistance, nutrient  
377 competition, allelopathy), we do not understand the relative importance of these elements during  
378 all stages of a bloom cycle. To this end, co-culturing of HAB and non-HAB species (plus grazers)  
379 needs to become more prevalent. While many studies (e.g. Chang and McClean, 1997; Cooper  
380 et al., 2016; Gallardo Rodríguez et al., 2009; Guerrini et al., 2007; Ignatiades et al., 2007; John  
381 and Flynn, 2000; Nascimento et al., 2005; Peperzak, 2003; Sala-Pérez et al., 2016; Varkitzi et al.,  
382 2010; Wang et al., 2014; Zhengbin et al., 2006) have investigated the physiological responses of  
383 individual HAB species to environmental conditions, only a few (e.g. Ji et al., 2011; Li et al., 2012;  
384 Poulin et al., 2018; Riegman et al., 1996; Wang and Tang, 2008) have added environmental

385 variability when looking at interactions between two or more species. Here, we used co-cultures  
386 to investigate how naturally co-occurring dinoflagellates are affected by changes in macronutrient  
387 availability and illustrate how consumer-resource models can be used to predict resource  
388 competition between multiple species in mixed batch cultures. This study demonstrates that  
389 consumer-resource modelling is a viable trait-based approach to understanding the dynamics of  
390 multiple species in mixed communities.

391

#### 392 4.1 Growth and competition

393 Two large batch culture experiments, for a combined total of 294 single and mixed cultures of five  
394 common dinoflagellates (*Alexandrium minutum*, *Prorocentrum lima*, *P. micans*, *Protoceratium*  
395 *reticulatum* and *Scrippsiella trochoidea*) spread across different nutrient regimes, were set up to  
396 explore whether consumer-resource modelling provides a good basis to understand interspecific  
397 interactions between dinoflagellates. As all the monoculture growth rates fell within the ranges  
398 expected from literature (Chang and McClean, 1997; Guerrini et al., 2007; Ignatiades et al., 2007;  
399 Lee et al., 2005; Nascimento et al., 2005; Peperzak, 2003; Sala-Pérez et al., 2016; Varkitzi et al.,  
400 2010; Wang et al., 2014), the growth rates reported here (ref. section 3.1 and Table 2) were  
401 considered representative for batch culture experiments. Modifications of the (macro)nutrient  
402 concentration in growth media are commonly used to study the effect of nutrient availability and  
403 stoichiometry on the growth of dinoflagellates (e.g. Cooper et al., 2016; Gallardo Rodríguez et al.,  
404 2009; Guerrini et al., 2007; Varkitzi et al., 2010; Zhengbin et al., 2006). This study occasionally  
405 found differences in growth rates between N:P ratios (experiment 1), but no linear or unimodal  
406 relationships were detected. Given that other studies have failed to find a relation between the  
407 growth rates of dinoflagellates and the relative availability of nutrients (John and Flynn, 2000; Li  
408 et al., 2012; Rhee, 1978; Varkitzi et al., 2010), the differences found here might be caused by  
409 intraspecific variation. It should, however, be noted that a small range of resource ratios was used,  
410 and that far more extreme N:P ratios are found in natural environments. Whether or not extreme

411 N:P ratios have significant effects on the growth of dinoflagellates cannot be deduced from our  
412 results. The increase in growth rate between orders of magnitude of nutrient availability (i.e.  
413 CF100 and CF10) should also be interpreted cautiously; the CF10 growth rates may have been  
414 overestimated due to the lower number of time points between the lag phase and the stationary  
415 phase for these treatments.

416 Shifts in growth rates caused by changes in either the relative (N:P) or the absolute (CF) nutrient  
417 concentrations did not change the dinoflagellates' community structure; *P. micans* attained the  
418 highest growth rate in all cultures. As interspecific competition in discontinuous cultures tends to  
419 favour whichever species grows fastest under the conditions used (Riegman et al., 1996), it is  
420 normal that *P. micans* dominated all mixed cultures. According to the mean parameter estimates  
421 of our consumer-resource model (CRM), the success of *P. micans* should be attributed to its  
422 ability to capture resources rather than a high resource efficiency or low natural mortality rates.  
423 The uptake probability of both nitrogen and phosphorus of *P. micans* were (among) the highest  
424 observed. All pelagic dinoflagellates grew at roughly the same rate relative to their monocultures  
425 in the early stages of both experiments. By sequestering nitrogen and phosphorus more rapidly,  
426 thereby denying its competitors access to these nutrients, *P. micans* was able to outgrow all other  
427 species in mixed cultures. Conversely, the benthic dinoflagellate *P. lima* was able to significantly  
428 increase its growth rate in mixed cultures. The difference in growth characteristics between its  
429 monocultures and the mixed cultures might have been caused by the release of organic nutrients  
430 by decaying cells of pelagic competitors. Sahraoui et al. (2013) have proposed that the growth of  
431 *P. lima* inside a lagoon can be triggered by organic matter, but little is known about the growth of  
432 this species on organic substances. Another unknown is whether the success of *P. micans* can  
433 be attributed to "luxury consumption". The rapid acquisition and storage of excess nutrients may  
434 be used to pre-emptively reduce the availability of resources for competing species (Droop, 1973;  
435 de Mazancourt & Schwartz, 2012). This trait has not been studied in *P. micans* to our knowledge,

436 but its carrying capacity is known to positively correlate with nitrogen concentrations (Zhengbin et  
437 al., 2006; Zheng-fang et al., 1995). Similar results were found in this study.

438

#### 439 4.2 CRM: use and considerations

440 The results of our experiments should not be viewed as ecological stoichiometry research; testing  
441 the effect of nutrient stoichiometry on the growth of dinoflagellates requires the use of continuous  
442 cultures with controlled dilution rates (cfr. van de Waal et al., 2014). Only chemostats can be used  
443 to determine the resource requirements of each species at the same net population growth rate.  
444 In batch culture, the relative availability of external nutrients will rapidly change over the course  
445 of the experiment, thus altering the intended treatment. In this study, the N:P ratios and the CF's  
446 were merely used to introduce variability in the nitrate concentrations, which we then chose as  
447 the driver of the consumer-resource model used. We set out to determine the efficacy of consumer  
448 resource modelling. Starting with the simplest setup available, which is the batch culture, we found  
449 that CRMs could be used to predict species dominance resulting from interspecific competition  
450 between dinoflagellates in mixed cultures. CRMs can approximate the densities of both winning  
451 and losing algal species up to the plateau phase with a high degree of accuracy. Stark changes  
452 in growth rate between monocultures and multispecies cultures such as those observed in *P. lima*  
453 can, however, lead to poor predictions if the underlying mechanism is not fully understood and  
454 included in the structural equations. By using a CRM, this study was able to demonstrate that the  
455 presence of a fast-growing species (*P. micans*) had strong, indirect negative effects on the growth  
456 of competing dinoflagellates; the growth of competing algae was to a large degree hampered by  
457 diminishing nutrient availability due to uptake by *P. micans*.

458 All dinoflagellates used here may produce allelochemicals that affect algal growth in one way or  
459 another (Arzul et al., 1999; Fistarol et al., 2004; Ji et al., 2011; Sala-Pérez et al., 2016; Wang and  
460 Tang, 2008; Yang et al., 2008), but we did not explicitly test our strains ability to do so here. By  
461 using a CRM, we managed to accurately predict the community dynamics throughout the growth

462 phase of each mixed culture using the nutrient uptake rates, conversion efficiencies and natural  
 463 mortality rates of each species (Fig. 3). That is not to say that allelopathic interactions could not  
 464 have occurred here. More likely than not, nutrient stress coupled to higher cell densities caused  
 465 increasingly significant allelopathic interactions by the end of the experiments but, as it stands,  
 466 the prototypical CRM cannot mimic quiescence and transient community dynamics. For starters,  
 467 the model is prone to underestimate maximum densities as the predicted cellular growth is  
 468 coupled to external nutrient concentrations and, hence, stops once nutrients are depleted. In  
 469 reality, cell growth is based on internal nutrient concentrations (Droop, 1974), thus allowing  
 470 population growth to continue in the absence of external nutrients. A common solution is to use  
 471 cell-based nutrient quota to establish relationships between the growth rate, internal nutrient  
 472 reserves, and external resource availability (Flynn, 2008b). Yet, while Droop's cell-quota model  
 473 (1974) is a good descriptor of growth in laboratory cultures, it is not well suited for competition  
 474 modelling due to the need to distinguish cell quota per species (in addition to other concerns; see  
 475 Flynn, 2008b). An alternative solution could be to add a discrete time lag ( $\epsilon$ ) to the growth and  
 476 external nutrient relation (cfr. the delayed allelopathic interactions of Mukhopadhyay et al., 1998).  
 477 The time lag ( $\epsilon$ ) of each species should correspond to the difference between its time of peak  
 478 density and the time of nutrient depletion. Population growth would then be described by:

$$479 \quad (10) \frac{dX_i}{dt} = X_i \cdot \epsilon - M_i$$

$$480 \quad \text{with } \epsilon = t_{X_{max}} - t_{[NO_3^-]_{min}}$$

481 In addition to this time lag, the inclusion of allelopathy would likely improve the predictions beyond  
 482 the growth phase. In the current model, population decline can only occur as a result of natural  
 483 mortality as observed in monoculture. However, as shown here, the population decline in mixed  
 484 cultures is far steeper than in monoculture (supporting figures SF5 and SF12), resulting in poor  
 485 predictions of species abundance after some time. Interspecific interactions – be it allelopathy or  
 486 mixotrophy – can be added to the model by introducing density-dependent parameters. Similar to

487 the work on Lotka-Volterra models (cfr. Ji et al., 2011; Qiu et al., 2012; Tameishi et al., 2009;  
488 Wang et al., 2013), the CRM could be modified as follows (equation 11). This hybrid model would  
489 bring together both direct and indirect interactions between competing microalgae.

$$490 \quad (11) \frac{dX_i}{dt} = X_i \cdot \varepsilon - M_i - \sum_{j=1}^n \alpha_{ij} X_j$$

491 With  $\alpha_{ij}$  being the coefficient of interaction between species  $i$  and species  $j$ .

492

493 Note that this approach assumes both constant and linear relationships between the densities of  
494 the allelopathic species and allelopathic interactions, which is an oversimplification given that the  
495 excretion of allelochemicals and their effects are heavily context-dependent (Poulson et al., 2010).  
496 The approach will also capture, confound or conceal other interactions (e.g. mixotrophy, induced  
497 cyst formation, stimulatory interactions, pH change) if these facets are not specifically measured.  
498 Unfortunately, the data generated here are not well suited to test this proposed model. In order to  
499 determine the interaction between two species, the experimental design should include bi-algal  
500 cultures of all competitors and account for the aforementioned pitfalls. Regardless, CRMs could  
501 become key instruments for understanding various species-species interactions in HAB ecology,  
502 and should be developed further to that end. Given the success of the Maynard-Smith function  
503 (e.g. Bandyopadhyay, 2006; Chattopadhyay, 1996; Mandal et al., 2014; Mukhopadhyay et al.,  
504 2003, 1998; Solé et al., 2005), we believe that CRMs and hybrid models still hold a great, mostly  
505 unexplored potential to improve our understanding of HAB dynamics. Even if CRM-based in situ  
506 modelling proves ineffective, simple CRMs should become a staple analysis when conducting  
507 multispecies lab experiments; they provide enhanced insights in competition dynamics with  
508 minimal data requirements. Going forward, it is recommended that the findings are tested further  
509 by applying the basic CRM to comparable datasets, that the model improvements suggested here  
510 (time delay, allelopathy, or others) are explored using fit-for-purpose experimental designs, that

511 the virtue of CRMs to understand continuous multispecies cultures (incl. grazing) is explored, and  
512 that additional nutrient sources (e.g. Si) are reintroduced into the model.

513

## 514 **5. Conclusions**

515 Consumer-resource modelling is a simple trait-based approach that has been used to understand  
516 coexistence dynamics in fields ranging from plant ecology to oncology. To date, however, CRMs  
517 are not commonly used in HAB research. This study shows that consumer-resource models can  
518 be used on the most common growth setup – the batch culture - with minimal data requirements,  
519 and that they provide key benefits to understanding resource competition between dinoflagellates.  
520 Based on our results and the success of Lotka-Volterra-based modelling approaches, we believe  
521 that the application of CRMs and derivatives should be explored further, both as a lab-tool as well  
522 as for in situ HAB modelling.

523

## 524 **Author contributions**

525 M.V. and C.J. acquired the main funding for this study. M.D.R. and C.J. designed the experiments.  
526 J.B. and F.D.L. provided the framework of the consumer-resource model. M.D.R. carried out the  
527 experiments and implemented the CRM with the help of J.B. N.B. performed the nutrient analyses.  
528 M.D.R. wrote the manuscript with input from all authors.

529

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537

## 538 **Supporting Figures**

539 **SF1-SF4:** Monoculture growth of *P. lima*, *P. micans*, *P. reticulatum*, and *S. trochoidea* at various  
540 N:P ratios (experiment 1) fitted with exponential or logistic growth models.

541 **SF5:** Growth of *P. lima*, *P. micans*, *P. reticulatum*, and *S. trochoidea* in multispecies cultures at  
542 various N:P ratios (experiment 1).

543 **SF6-SF9:** Monoculture growth of *P. lima*, *P. micans*, *P. reticulatum*, and *S. trochoidea* at various  
544 N:P ratios (experiment 1), densities and nutrients fitted with a consumer-resource model.

545 **SF10:** Growth of *P. lima*, *P. micans*, *P. reticulatum*, and *S. trochoidea* in mixed cultures at various  
546 N:P ratios (experiment 1): nutrients and densities of each species fitted with a consumer-resource  
547 model based on parameter estimates from monocultures.

548 **SF11:** Monoculture growth of *A. minutum*, *P. micans* and *P. reticulatum* at three N:P ratios and  
549 four concentration factors (CFs). Black = CF100; Blue = CF10. Data from the two highest CFs  
550 was fitted with logistic growth models.

551 **SF12:** Growth of *A. minutum* (red), *P. micans* (black) and *P. reticulatum* (blue) in mixed cultures  
552 at three N:P ratios and four concentration factors (CFs): Data of *P. micans* was fitted with logistic  
553 growth models.

554 **SF13-15:** Monoculture growth of *A. minutum*, *P. micans* and *P. reticulatum* in mixed cultures at  
555 three N:P ratios and two concentration factors (CFs) fitted with a consumer-resource model.

556 **SF16:** Growth of *A. minutum* (red), *P. micans* (black) and *P. reticulatum* (blue) in mixed cultures  
557 at three N:P ratios and two concentration factors (CFs): nutrients and densities of each species  
558 fitted with a consumer-resource model based on parameter estimates from monocultures.

559 Table 1: Mean growth rates and carrying capacities of *A. minutum*, *P. reticulatum* and *P. micans*  
 560 grown in either single or mixed cultures at different N:P ratios and two concentration factors (CF),  
 561 representing 100% or 10% v/v dilutions of L1 medium and artificial seawater at a N:P ratio of 24.  
 562 Results shown are from the second experiment. Values represent  $\mu \pm s.d.$

563

Species	N:P	CF	$\mu_{mono}$ (d <sup>-1</sup> )	$K_{mono}$ (10 <sup>8</sup> $\mu\text{m}^3 \cdot \text{ml}^{-1}$ )	$\mu_{mix}$ (d <sup>-1</sup> )	$K_{mix}$ (10 <sup>8</sup> $\mu\text{m}^3 \cdot \text{ml}^{-1}$ )
<i>A. minutum</i>	8	100	0.28±0.01	7.09±0.15	0.25±0.01	1.03±0.26
		10	0.31±0.02	0.99±0.06	0.35±0.02	0.13±0.01
	16	100	0.25±0.00	12.3±0.56	0.24±0.09	0.77±0.22
		10	0.32±0.02	1.36±0.01	0.36±0.01	0.13±0.02
	24	100	0.25±0.01	15.6±0.40	0.23±0.09	0.81±0.17
		10	0.33±0.02	1.59±0.05	0.35±0.02	0.19±0.02
<i>P. micans</i>	8	100	0.28±0.01	4.73±0.33	0.26±0.03	4.10±0.77
		10	0.36±0.01	0.67±0.04	0.33±0.04	0.52±0.07
	16	100	0.28±0.02	6.52±0.73	0.24±0.03	5.61±0.03
		10	0.38±0.02	0.88±0.06	0.33±0.00	0.67±0.02
	24	100	0.30±0.00	6.86±0.02	0.27±0.01	5.25±0.28
		10	0.37±0.02	1.02±0.02	0.30±0.04	0.69±0.04
<i>P. reticulatum</i>	8	100	0.18±0.01	3.29±0.21	0.25±0.12	0.27±0.20
		10	0.21±0.03	0.47±0.01	0.28±0.06	0.08±0.07
	16	100	0.17±0.01	6.75±0.10	0.16±0.09	0.71±0.40
		10	0.19±0.02	0.65±0.04	0.28±0.06	0.07±0.00
	24	100	0.19±0.01	7.76±0.20	0.16±0.03	0.74±0.11
		10	0.28±0.03	0.48±0.02	0.15±0.03	0.12±0.07

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565

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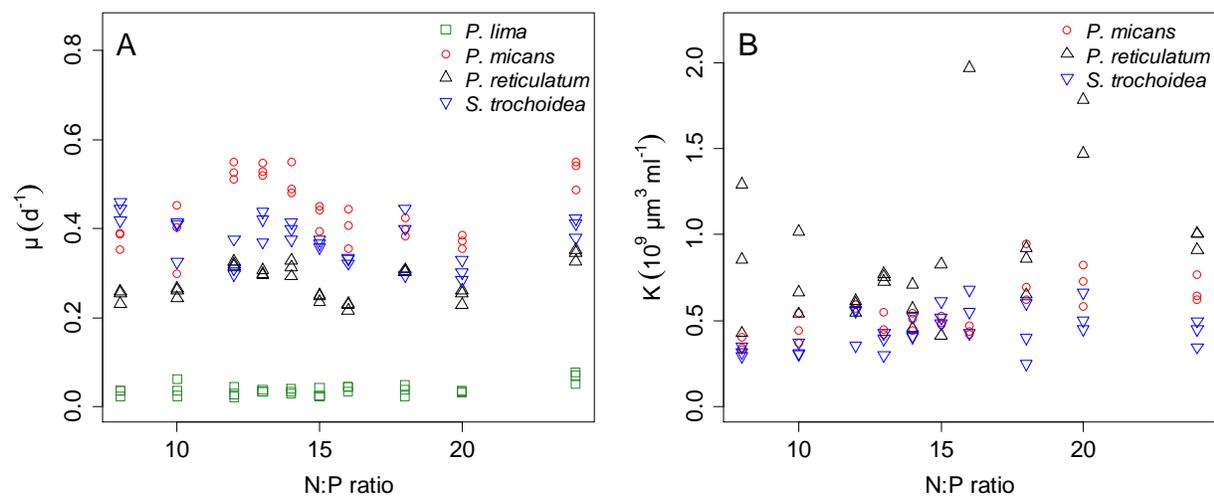
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568

569 Table 2: Mean parameter estimates derived from monocultures and used to predict cell growth in  
 570 multispecies cultures with a simplified version of MacArthur's consumer-resource model (1970).  
 571 The results are calculated based on a 1000 Monte-Carlo simulations, each randomly drawing  
 572 from the prior distributions generated by Markov chain Monte Carlo (MCMC) simulations.  $U_{NO_3}$  is  
 573 the uptake probability of  $NO_3$  per unit biovolume of a dinoflagellate per time unit;  $U_{PO_4}$  is the uptake  
 574 probability of  $PO_4$  per unit biovolume of a dinoflagellate per time unit;  $W_{NO_3}$  is the efficiency at  
 575 which nitrogen is converted into biovolume;  $M$  is the fraction of biovolume lost daily due to natural  
 576 mortality. Values shown are the averages ( $\pm$ s.d.) per experiment (Exp) across all N:P ratios.

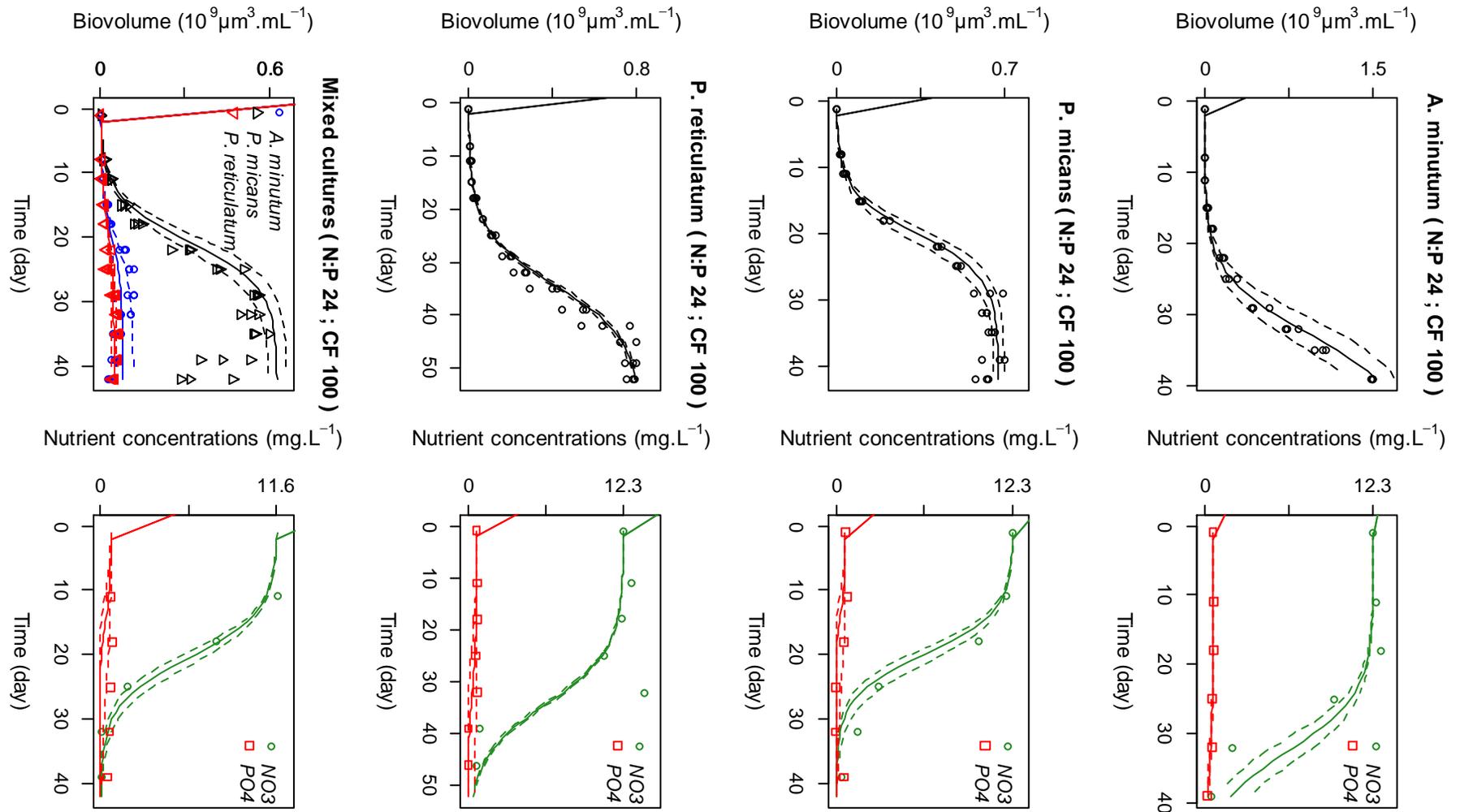
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<b>Species</b>	<b>Exp</b>	<b>CF</b>	<b><math>U_{NO_3}</math></b> ( $10^{-10} \mu m^{-3} \cdot d^{-1}$ )	<b><math>U_{PO_4}</math></b> ( $10^{-10} \mu m^{-3} \cdot d^{-1}$ )	<b><math>W_{NO_3}</math></b> ( $\mu m^3 \cdot pg^{-1}$ )	<b><math>M</math></b> ( $10^{-6} d^{-1}$ )
<i>P. lima</i>	1	100	1.7 $\pm$ 3.1	8.2 $\pm$ 3.9	0.26 $\pm$ 0.38	4.9 $\pm$ 0.6
<i>P. micans</i>	1	100	11 $\pm$ 3.5	7.9 $\pm$ 11	0.06 $\pm$ 0.01	4.6 $\pm$ 0.7
<i>P. reticulatum</i>	1	100	3.7 $\pm$ 1.4	2.6 $\pm$ 1.8	0.10 $\pm$ 0.04	5.2 $\pm$ 0.4
<i>S. trochoidea</i>	1	100	13 $\pm$ 5.6	2.9 $\pm$ 7.7	0.04 $\pm$ 0.01	5.1 $\pm$ 0.7
<i>A. minutum</i>	2	100	2.2 $\pm$ 1.0	0.3 $\pm$ 0.3	0.17 $\pm$ 0.02	5.2 $\pm$ 0.4
	2	10	26 $\pm$ 5.8	32 $\pm$ 22	0.18 $\pm$ 0.05	5.4 $\pm$ 0.8
<i>P. micans</i>	2	100	5.2 $\pm$ 1.1	6.2 $\pm$ 9.9	0.08 $\pm$ 0.02	6.1 $\pm$ 0.9
	2	10	49 $\pm$ 9.9	11 $\pm$ 6.2	0.11 $\pm$ 0.03	4.6 $\pm$ 0.6
<i>P. reticulatum</i>	2	100	3.5 $\pm$ 1.8	2.3 $\pm$ 1.4	0.08 $\pm$ 0.01	5.0 $\pm$ 1.4
	2	10	42 $\pm$ 9.2	59 $\pm$ 9.8	0.08 $\pm$ 0.03	5.7 $\pm$ 0.6



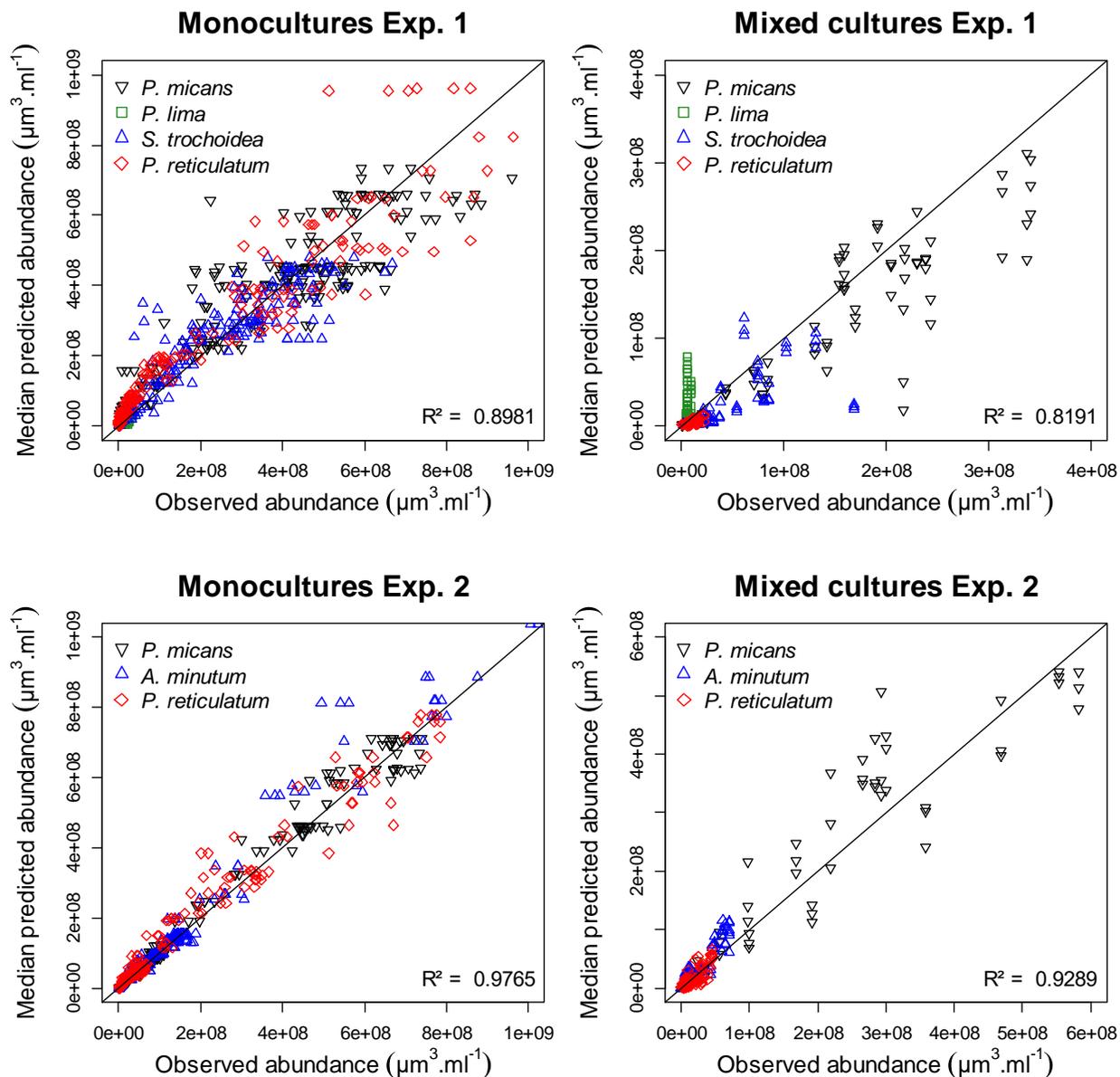
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579 Fig. 1: (A) growth rates per monoculture of *P. micans* (red), *P. lima* (green), *P. reticulatum* (black)580 and *S. trochoidea* (blue) across different nitrogen-to-phosphorus ratio's; (B) carrying capacities of581 monocultures of *P. micans* (red), *P. reticulatum* (black) and *S. trochoidea* (blue) across N:P ratio's.582 All results were obtained from the first experiment. All cultures, except those of *P. lima*, were fitted583 with logistic growth models. *P. lima* was fitted with exponential growth models.



584

585 Fig. 2: Monoculture data of *A. minutum*, *P. micans*, and *P. reticulatum* was used to parametrize a consumer-resource model and predict  
 586 the growth of each dinoflagellate in mixed cultures. The example shown here is from the second experiment, using regular L1 medium  
 587 (N:P 24; CF100). Full lines are the average predicted abundance of a 1000 Monte Carlo simulations, randomly drawing from posterior  
 588 parameter distributions made with Markov Chain Monte Carlo methods following simulated annealing. The dotted lines represent the  
 589 5%-95% confidence interval around these averages. Markers are data as observed.



590

591 Fig. 3: Goodness-of-fit of a simplified consumer-resource model of MacArthur (1970), applied to

592 biovolumes from monocultures (left) and multispecies cultures (right) of two growth experiments.

593 Data shown reflect predicted vs. observed abundances up to and including the plateau-phase.

594 **6. References**

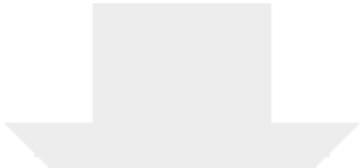
- 595 Abrams, P., 1975. Limiting similarity and the form of the competition coefficient. *Theor. Popul. Biol.* 8,  
596 356–375.
- 597 Allen, J.I., Polimene, L., 2011. Linking physiology to ecology: towards a new generation of plankton  
598 models. *J. Plankton Res.* 33, 989–997.
- 599 Allen, J.L., Ten-Hage, L., Leflaive, J., 2016. Allelopathic interactions involving benthic phototrophic  
600 microorganisms. *Environ. Microbiol. Rep.* 8, 752–762.
- 601 Anderson, D.M., Glibert, P.M., Burkholder, J.M., 2002. Harmful algal blooms and eutrophication:  
602 Nutrient sources, composition, and consequences. *Estuaries* 25, 704–726.
- 603 Armstrong, R.A., 1979. Prey Species Replacement along a Gradient of Nutrient Enrichment: A Graphical  
604 Approach. *Ecology* 60, 76–84.
- 605 Arzul, G., Seguel, M., Guzman, L., Erard-Le Denn, E., 1999. Comparison of allelopathic properties in three  
606 toxic *Alexandrium* species. *J. Exp. Mar. Biol. Ecol.* 232, 285–295.
- 607 Balch, W.M., 2004. Re-evaluation of the physiological ecology of coccolithophores, in: Thierstein,  
608 P.D.H.R., Young, D.J.R. (Eds.), *Coccolithophores*. Springer Berlin Heidelberg, pp. 165–190.
- 609 Bandyopadhyay, M., 2006. Dynamical analysis of an allelopathic phytoplankton model. *J. Biol. Syst.* 14,  
610 205–217.
- 611 Baty, F., Ritz, C., Charles, S., Brutsche, M., Flandrois, J.-P., Delignette-Muller, M.-L., 2015. A Toolbox for  
612 Nonlinear Regression in R: The Package nlstools. *J. Stat. Softw.* 66, 1–21.
- 613 Blossom, H.E., Markussen, B., Daugbjerg, N., Krock, B., Norlin, A., Hansen, P.J., 2019. The Cost of Toxicity  
614 in Microalgae: Direct Evidence From the Dinoflagellate *Alexandrium*. *Front. Microbiol.* 10.
- 615 Bravo, I., Figueroa, R., 2014. Towards an Ecological Understanding of Dinoflagellate Cyst Functions.  
616 *Microorganisms* 2, 11–32.
- 617 Chakraborty, S., Ramesh, A., Dutta, P.S., 2015. Toxic phytoplankton as a keystone species in aquatic  
618 ecosystems: stable coexistence to biodiversity. *Oikos* 5, 735–746.
- 619 Chang, F.H., McClean, M., 1997. Growth responses of *Alexandrium minutum* (Dinophyceae) as a function  
620 of three different nitrogen sources and irradiance. *N. Z. J. Mar. Freshw. Res.* 31, 1–7.
- 621 Chattopadhyay, J., 1996. Effect of toxic substances on a two-species competitive system. *Ecol. Model.*  
622 84, 287–289.
- 623 Chesson, P., 1990. MacArthur's consumer-resource model. *Theor. Popul. Biol.* 37, 26–38.
- 624 Cooper, J.T., Sinclair, G.A., Wawrik, B., 2016. Transcriptome Analysis of *Scrippsiella trochoidea* CCMP  
625 3099 Reveals Physiological Changes Related to Nitrate Depletion. *Front. Microbiol.* 7, 639.
- 626 Crane, K.W., Grover, J.P., 2010. Coexistence of mixotrophs, autotrophs, and heterotrophs in planktonic  
627 microbial communities. *J. Theor. Biol.* 262, 517–527.
- 628 Dam, H.G., Haley, S.T., 2011. Comparative dynamics of paralytic shellfish toxins (PST) in a tolerant and  
629 susceptible population of the copepod *Acartia hudsonica*. *Harmful Algae* 10, 245–253.
- 630 de Mazancourt, C., Schwartz, M.W., 2012. Starve a competitor: evolution of luxury consumption as a  
631 competitive strategy. *Theor. Ecol.* 5, 37–49.
- 632 Driscoll, W.W., Hackett, J.D., Ferrière, R., 2016. Eco-evolutionary feedbacks between private and public  
633 goods: evidence from toxic algal blooms. *Ecol. Lett.* 19, 81–97.
- 634 Droop, M.R., 1973. Some Thoughts on Nutrient Limitation in Algae. *J. Phycol.* 9, 264–272.
- 635 Droop, M.R., 1974. The nutrient status of algal cells in continuous culture. *J. Mar. Biol. Assoc. U. K.* 54,  
636 825–855.
- 637 Dunn, O.J., 1964. Multiple Comparisons Using Rank Sums. *Technometrics* 6, 241–252.
- 638 Eppley, R.W., 1972. Temperature and Phytoplankton Growth in Sea. *Fish. Bull. Natl. Ocean. Atmospheric*  
639 *Adm.* 70, 1063–1085.

- 640 Fistarol, G.O., Legrand, C., Selander, E., Hummert, C., Stolte, W., Granéli, E., 2004. Allelopathy in  
641 *Alexandrium* spp.: effect on a natural plankton community and on algal monocultures. *Aquat.*  
642 *Microb. Ecol.* 35, 45–56.
- 643 Flynn, K.J., 2008a. Attack is not the best form of defense: Lessons from harmful algal bloom dynamics.  
644 *Harmful Algae* 8, 129–139.
- 645 Flynn, K.J., 2008b. Use, abuse, misconceptions and insights from quota models — the Droop cell quota  
646 model 40 years on, in: *Oceanography and Marine Biology: An Annual Review*. pp. 1–23.
- 647 Gallardo Rodríguez, J.J., Sánchez Mirón, A., Cerón García, M. del C., Belarbi, E.H., García Camacho, F.,  
648 Chisti, Y., Molina Grima, E., 2009. Macronutrients requirements of the dinoflagellate  
649 *Protoceratium reticulatum*. *Harmful Algae* 8, 239–246.
- 650 Glibert, P.M., 2016. Margalef revisited: A new phytoplankton mandala incorporating twelve dimensions,  
651 including nutritional physiology. *Harmful Algae* 55, 25–30.
- 652 Glibert, P.M., Allen, J.I., Bouwman, A.F., Brown, C.W., Flynn, K.J., Lewitus, A.J., Madden, C.J., 2010.  
653 Modeling of HABs and eutrophication: Status, advances, challenges. *J. Mar. Syst., GEOHAB*  
654 *Modeling* 83, 262–275.
- 655 Glibert, P.M., Burkholder, J.M., 2006. The Complex Relationships Between Increases in Fertilization of  
656 the Earth, Coastal Eutrophication and Proliferation of Harmful Algal Blooms, in: Granéli, E.,  
657 Turner, J.T. (Eds.), *Ecology of Harmful Algae*, Ecological Studies. Springer, Berlin, Heidelberg, pp.  
658 341–354.
- 659 Granéli, E., Hansen, P.J., 2006. Allelopathy in Harmful Algae: A Mechanism to Compete for Resources?,  
660 in: Granéli, P.D.E., Turner, P.D.J.T. (Eds.), *Ecology of Harmful Algae*, Ecological Studies. Springer  
661 Berlin Heidelberg, pp. 189–201.
- 662 Granéli, E., Salomon, P.S., Fistarol, G.O., 2008a. The Role of Allelopathy for Harmful Algae Bloom  
663 Formation, in: Evangelista, V., Barsanti, L., Frassanito, A.M., Passarelli, V., Gualtieri, P. (Eds.),  
664 *Algal Toxins: Nature, Occurrence, Effect and Detection*, NATO Science for Peace and Security  
665 Series A: Chemistry and Biology. Springer Netherlands, pp. 159–178.
- 666 Granéli, E., Weberg, M., Salomon, P.S., 2008b. Harmful algal blooms of allelopathic microalgal species:  
667 The role of eutrophication. *Harmful Algae* 8, 94–102.
- 668 Gröger, M., Maier-Reimer, E., Mikolajewicz, U., Moll, A., Sein, D., 2013. NW European shelf under  
669 climate warming: implications for open ocean - shelf exchange, primary production, and carbon  
670 absorption. *Biogeosciences* 10, 3767–3792.
- 671 Guerrini, F., Ciminiello, P., Dell’Aversano, C., Tartaglione, L., Fattorusso, E., Boni, L., Pistocchi, R., 2007.  
672 Influence of temperature, salinity and nutrient limitation on yessotoxin production and release  
673 by the dinoflagellate *Protoceratium reticulatum* in batch-cultures. *Harmful Algae* 6, 707–717.
- 674 Guillard, R.R.L., Hargraves, P.E., 1993. *Stichochrysis immobilis* is a diatom, not a chrysophyte. *Phycologia*  
675 32, 234–236.
- 676 Hansen, H.P., Koroleff, F., 1999. Determination of nutrients, in: Grasshoff, K., Kremling, K., Ehrhardt, M.,  
677 (Eds.), *Methods of Seawater Analysis*. Wiley-VCH Verlag GmbH, pp. 159–228.
- 678 Hastings, W.K., 1970. Monte Carlo Sampling Methods Using Markov Chains and Their Applications.  
679 *Biometrika* 57, 97–109.
- 680 Heisler, J., Glibert, P.M., Burkholder, J.M., Anderson, D.M., Cochlan, W., Dennison, W.C., Dortch, Q.,  
681 Gobler, C.J., Heil, C.A., Humphries, E., Lewitus, A., Magnien, R., Marshall, H.G., Sellner, K.,  
682 Stockwell, D.A., Stoecker, D.K., Suddleson, M., 2008. Eutrophication and harmful algal blooms: A  
683 scientific consensus. *Harmful Algae, HABs and Eutrophication* 8, 3–13.
- 684 Huisman, J., Weissing, F.J., 1994. Light-Limited Growth and Competition for Light in Well-Mixed Aquatic  
685 Environments: An Elementary Model. *Ecology* 75, 507–520.

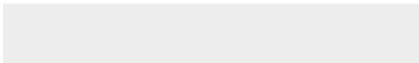
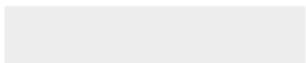
- 686 Ianora, A., Bentley, M.G., Caldwell, G.S., Casotti, R., Cembella, A.D., Engström-Öst, J., Halsband, C.,  
687 Sonnenschein, E., Legrand, C., Llewellyn, C.A., Paldavičienė, A., Pilkaitė, R., Pohnert, G.,  
688 Razinkovas, A., Romano, G., Tillmann, U., Vaiciute, D., 2011. The Relevance of Marine Chemical  
689 Ecology to Plankton and Ecosystem Function: An Emerging Field. *Mar. Drugs* 9, 1625–1648.
- 690 Ignatiades, L., Gotsis-Skretas, O., Metaxatos, A., 2007. Field and culture studies on the ecophysiology of  
691 the toxic dinoflagellate *Alexandrium minutum* (Halim) present in Greek coastal waters. *Harmful*  
692 *Algae* 6, 153–165.
- 693 Ji, X., Han, X., Zheng, L., Yang, B., Yu, Z., Zou, J., 2011. Allelopathic interactions between *Prorocentrum*  
694 *micans* and *Skeletonema costatum* or *Karenia mikimotoi* in laboratory cultures. *Chin. J. Oceanol.*  
695 *Limnol.* 29, 840–848.
- 696 John, E.H., Flynn, K.J., 2000. Growth dynamics and toxicity of *Alexandrium fundyense* (Dinophyceae): the  
697 effect of changing N:P supply ratios on internal toxin and nutrient levels. *Eur. J. Phycol.* 35, 11–  
698 23.
- 699 Jonsson, P.R., Pavia, H., Toth, G., 2009. Formation of harmful algal blooms cannot be explained by  
700 allelopathic interactions. *Proc. Natl. Acad. Sci. U. S. A.* 106, 11177–11182.
- 701 Kruskal, W.H., Wallis, W.A., 1952. Use of Ranks in One-Criterion Variance Analysis. *J. Am. Stat. Assoc.* 47,  
702 583–621.
- 703 Lee, C.-K., Lee, O.-H., Lee, S.-G., 2005. Impacts of Temperature, Salinity and Irradiance on the Growth of  
704 Ten Harmful Algal Bloom-forming Microalgae Isolated in Korean Coastal Waters. *The Sea* 10, 79–  
705 91.
- 706 Legrand, C., Rengefors, K., Fistarol, G.O., Granéli, E., 2003. Allelopathy in phytoplankton - biochemical,  
707 ecological and evolutionary aspects. *Phycologia* 42, 406–419.
- 708 Li, J., Glibert, P.M., Alexander, J.A., Molina, M.E., 2012. Growth and competition of several harmful  
709 dinoflagellates under different nutrient and light conditions. *Harmful Algae* 13, 112–125.
- 710 MacArthur, R., 1970. Species packing and competitive equilibrium for many species. *Theor. Popul. Biol.*  
711 1, 1–11.
- 712 MacArthur, R., 1969. Species packing, and what competition minimizes. *Proc. Natl. Acad. Sci. U. S. A.* 64,  
713 1369–1371.
- 714 MacArthur, R., Levins, R., 1967. The Limiting Similarity, Convergence, and Divergence of Coexisting  
715 Species. *Am. Nat.* 101, 377–385.
- 716 Mandal, P.S., Allen, L.J.S., Banerjee, M., 2014. Stochastic modeling of phytoplankton allelopathy. *Appl.*  
717 *Math. Model.* 38, 1583–1596.
- 718 Margalef, R., 1978. Life-forms of phytoplankton as survival alternatives in an unstable environment.  
719 *Oceanol. Acta* 1, 493–509.
- 720 Maynard Smith, J., 1974. *Models in Ecology*. Cambridge University Press, New York.
- 721 Mortelmans, J., Deneudt, K., Cattrijse, A., Beauchard, O., Daveloose, I., Vyverman, W., Vanaverbeke, J.,  
722 Timmermans, K., Peene, J., Roose, P., Knockaert, M., Chou, L., Sanders, R., Stinchcombe, M.,  
723 Kimpe, P., Lammens, S., Theetaert, H., Gkritzalis, T., Hernandez, F., Mees, J., 2019. Nutrient,  
724 pigment, suspended matter and turbidity measurements in the Belgian part of the North Sea.  
725 *Sci. Data* 6, 1–8.
- 726 Mukhopadhyay, A., Chattopadhyay, J., Tapaswi, P.K., 1998. A delay differential equations model of  
727 plankton allelopathy. *Math. Biosci.* 23.
- 728 Mukhopadhyay, A., Tapaswi, P.K., Chattopadhyay, J., 2003. A space-time state-space model of  
729 phytoplankton allelopathy. *Nonlinear Anal. Real World Appl.* 4, 437–456.
- 730 Nascimento, S.M., Purdie, D.A., Morris, S., 2005. Morphology, toxin composition and pigment content of  
731 *Prorocentrum lima* strains isolated from a coastal lagoon in southern UK. *Toxicon* 45, 633–649.
- 732 Peperzak, L., 2003. Climate change and harmful algal blooms in the North Sea. *Acta Oecologica* 24,  
733 S139–S144.

- 734 Poulin, R.X., Poulson-Ellestad, K.L., Roy, J.S., Kubanek, J., 2018. Variable allelopathy among  
735 phytoplankton reflected in red tide metabolome. *Harmful Algae* 71, 50–56.
- 736 Poulson, K., Sieg, R., Prince, E., Kubanek, J., 2010. Allelopathic compounds of a red tide dinoflagellate  
737 have species-specific and context-dependent impacts on phytoplankton. *Mar. Ecol. Prog. Ser.*  
738 416, 69–78.
- 739 Reigosa, M.J., Sánchez-Moreiras, A., González, L., 1999. Ecophysiological Approach in Allelopathy. *Crit.*  
740 *Rev. Plant Sci.* 18, 577–608.
- 741 Rhee, G.-Y., 1978. Effects of N:P atomic ratios and nitrate limitation on algal growth, cell composition,  
742 and nitrate uptake. *Limnol. Oceanogr.* 23, 10–25.
- 743 Richerson, P., Armstrong, R., Goldman, C.R., 1970. Contemporaneous Disequilibrium, a New Hypothesis  
744 to Explain the "Paradox of the Plankton". *Proc. Natl. Acad. Sci. U. S. A.* 67, 1710–1714.
- 745 Riegman, R., Boer, M. de, Domis, L. de S., 1996. Growth of harmful marine algae in multispecies cultures.  
746 *J. Plankton Res.* 18, 1851–1866.
- 747 Roy, S., Chattopadhyay, J., 2007. Toxin-allelopathy among phytoplankton species prevents competitive  
748 exclusion. *J. Biol. Syst.* 15, 73–93.
- 749 Sahraoui, I., Bouchouicha, D., Mabrouk, H.H., Hlaili, A.S., 2013. Driving factors of the potentially toxic  
750 and harmful species of *Prorocentrum Ehrenberg* in a semi-enclosed Mediterranean lagoon  
751 (Tunisia, SW Mediterranean). *Mediterr. Mar. Sci.* 14, 353–362.
- 752 Sala-Pérez, M., Alpermann, T.J., Krock, B., Tillmann, U., 2016. Growth and bioactive secondary  
753 metabolites of arctic *Protoceratium reticulatum* (Dinophyceae). *Harmful Algae* 55, 85–96.
- 754 Smayda, T.J., 2008. Complexity in the eutrophication–harmful algal bloom relationship, with comment  
755 on the importance of grazing. *Harmful Algae* 8, 140–151.
- 756 Smayda, T.J., 1997. Harmful algal blooms: Their ecophysiology and general relevance to phytoplankton  
757 blooms in the sea. *Limnol. Oceanogr.* 42, 1137–1153.
- 758 Solé, J., García-Ladona, E., Ruardij, P., Estrada, M., 2005. Modelling allelopathy among marine algae.  
759 *Ecol. Model.* 183, 373–384.
- 760 Sourisseau, M., Plus, M., Chapelle, A., Le Guennec, V., Le Gland, G., 2017. How competition for resources  
761 drive specific niches and community structure of phytoplankton by using a trait-based model.  
762 *Front. Mar. Sci.* 4.
- 763 Stoecker, D.K., Thessen, A.E., Gustafson, D.E., 2008. "Windows of opportunity" for dinoflagellate  
764 blooms: Reduced microzooplankton net growth coupled to eutrophication. *Harmful Algae* 8,  
765 158–166.
- 766 Sweeney, B.M., 1978. Opening remarks, session I: The organisms, in: Taylor, D.L., Seliger, H.H. (Eds.),  
767 Toxic Dinoflagellate Blooms. Proceedings of the 2nd International Conference on Toxic  
768 Dinoflagellate Blooms. Elsevier/North Holland, New York, pp. 37–40.
- 769 Sweeney, B.M., 1975. Red tides I have known., in: LoCicero, V.R. (Ed.), Toxic Dinoflagellate Blooms.  
770 Proceedings of the 1st International Conference on Toxic Dinoflagellate Blooms. Massachusetts  
771 Science & Technology Foundation, Wakefield, Massachusetts, pp. 225–234.
- 772 Tillmann, U., 2004. Interactions between planktonic microalgae and protozoan grazers. *J. Eukaryot.*  
773 *Microbiol.* 51, 156–168.
- 774 Tilman, D., 1977. Resource Competition between Plankton Algae: An Experimental and Theoretical  
775 Approach. *Ecology* 58, 338–348.
- 776 Turner, J.T., 2006. Harmful Algae Interactions with Marine Planktonic Grazers, in: Granéli, P.D.E., Turner,  
777 P.D.J.T. (Eds.), Ecology of Harmful Algae, Ecological Studies. Springer Berlin Heidelberg, pp. 259–  
778 270.
- 779 van de Waal, D.B., Eberlein, T., Bublitz, Y., John, U., Rost, B., 2014. Shake it easy: a gently mixed  
780 continuous culture system for dinoflagellates. *J. Plankton Res.* 36, 889–894.

- 781 Varkitzi, I., Pagou, K., Granéli, E., Hatzianestis, I., Pyrgaki, C., Pavlidou, A., Montesanto, B., Economou-  
782 Amilli, A., 2010. Unbalanced N:P ratios and nutrient stress controlling growth and toxin  
783 production of the harmful dinoflagellate *Prorocentrum lima* (Ehrenberg) Dodge. *Harmful Algae*  
784 9, 304–311.
- 785 Wang, Y., Tang, X., 2008. Interactions between *Prorocentrum donghaiense* Lu and *Scrippsiella trochoidea*  
786 (Stein) Loeblich III under laboratory culture. *Harmful Algae* 7, 65–75.
- 787 Wang, Z., Yu, Z., Song, X., Cao, X., Zhang, Y., 2014. Effects of ammonium and nitrate on encystment and  
788 growth of *Scrippsiella trochoidea*. *Chin. Sci. Bull.* 59, 4491–4497.
- 789 Weidenhamer, J.D., 2006. Distinguishing allelopathy from resource competition: the role of density, in:  
790 *Allelopathy*. Springer, pp. 85–103.
- 791 Wells, M.L., Trainer, V.L., Smayda, T.J., Karlson, B.S.O., Trick, C.G., Kudela, R.M., Ishikawa, A., Bernard, S.,  
792 Wulff, A., Anderson, D.M., Cochlan, W.P., 2015. Harmful algal blooms and climate change:  
793 Learning from the past and present to forecast the future. *Harmful Algae* 49, 68–93.
- 794 Wilkinson, G.N., Rogers, C.E., 1973. Symbolic Description of Factorial Models for Analysis of Variance. *J.*  
795 *R. Stat. Soc. Ser. C Appl. Stat.* 22, 392–399.
- 796 Xu, J., Kiørboe, T., 2018. Toxic dinoflagellates produce true grazer deterrents. *Ecology* 99, 2240–2249.
- 797 Yang, W., Li, L., Liu, J., Zhang, J., 2008. Allelopathy of marine benthic dinoflagellate *Prorocentrum lima* on  
798 three red tide algae. *Acta Sci. Circumstantiae* 28, 1631–1637.
- 799 Zhengbin, Z., Peifeng, L.I., Chunying, L., 2006. Effects of NO and different media on the growth of  
800 *Prorocentrum micans*. *J. Ocean Univ. China* 5, 239–242.
- 801 Zheng-fang, W., Qing, Z., Min, G., 1995. The effects of nitrogen, phosphorus, vitamins and trace metals  
802 on the growth of the red tide organism *Prorocentrum micans*. *Chin. J. Oceanol. Limnol.* 13, 338–  
803 342.

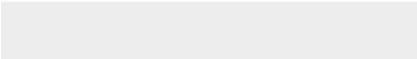
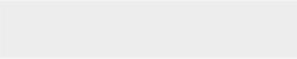


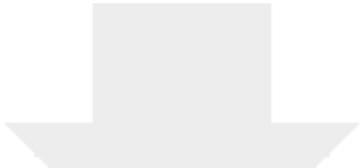
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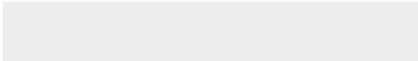
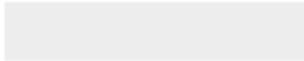


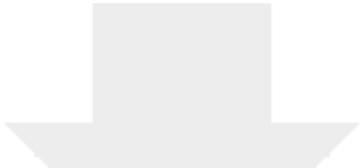
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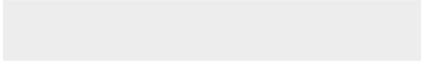
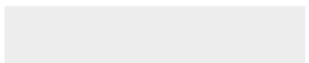


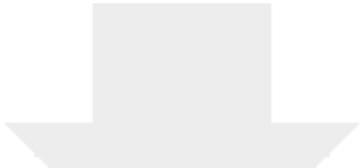
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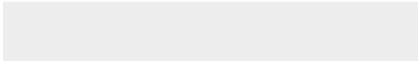
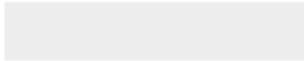


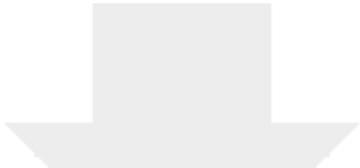
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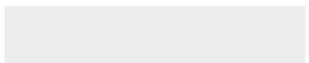


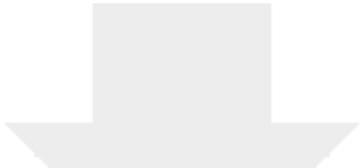
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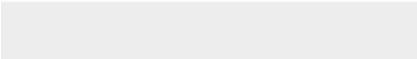
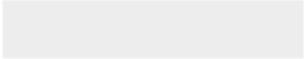


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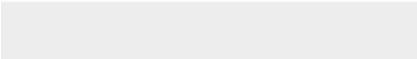
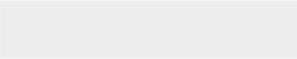


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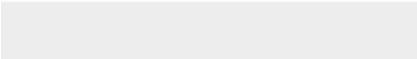
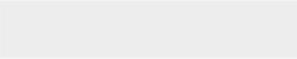


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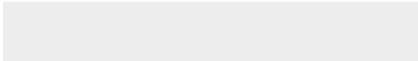
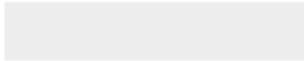


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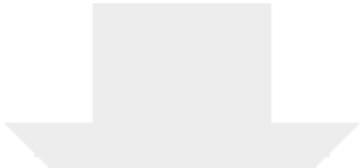


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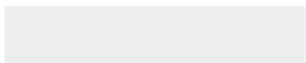




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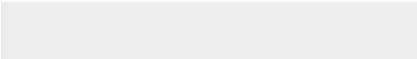
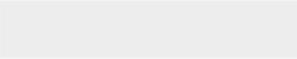


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## **HARMFUL ALGAE**

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